Modelling night-time ecosystem respiration by a constrained source optimization method

CHUN-TA LAI*, GABRIEL KATUL†, JOHN BUTNOR‡, DAVID ELLSWORTH§ and RAM OREN†

*Department of Biology, University of Utah, 2575. 1400E Salt Lake City, UT 84112-0840, USA; †Department of the Environment, Duke University, Durham NC 27708-0328, USA; ‡U.S.D.A. Forest Service, Southern Research Station, 3041 Cornwallis Road, Research Triangle Park, NC 27707, USA; § School of Natural Resources and the Environment, University of Michigan, Dana Building, 430 East University Avenue, Ann Arbor, MI 48109-1115, USA

Abstract

One of the main challenges to quantifying ecosystem carbon budgets is properly quantifying the magnitude of night-time ecosystem respiration. Inverse Lagrangian dispersion analysis provides a promising approach to addressing such a problem when measured mean CO2 concentration profiles and nocturnal velocity statistics are available. An inverse method, termed 'Constrained Source Optimization' or CSO, which couples a localized near-field theory (LNF) of turbulent dispersion to respiratory sources, is developed to estimate seasonal and annual components of ecosystem respiration. A key advantage to the proposed method is that the effects of variable leaf area density on flow statistics are explicitly resolved via higher-order closure principles. In CSO, the source distribution was computed after optimizing key physiological parameters to recover the measured mean concentration profile in a least-square fashion. The proposed method was field-tested using 1 year of 30-min mean CO2 concentration and CO2 flux measurements collected within a 17-year-old (in 1999) even-aged loblolly pine (Pinus taeda L.) stand in central North Carolina. Eddy-covariance flux measurements conditioned on large friction velocity, leaf-level porometry and forest-floor respiration chamber measurements were used to assess the performance of the CSO model. The CSO approach produced reasonable estimates of ecosystem respiration, which permits estimation of ecosystem gross primary production when combined with daytime net ecosystem exchange (NEE) measurements. We employed the CSO approach in modelling annual respiration of aboveground plant components (c. 214 g C m⁻² year⁻¹) and forest floor (c. 989 g C m⁻² year⁻¹) for estimating gross primary production (c. 1800 g C m⁻² year⁻¹) with a NEE of c. 605 g C m⁻² year⁻¹ for this pine forest ecosystem. We conclude that the CSO approach can utilise routine CO2 concentration profile measurements to corroborate forest carbon balance estimates from eddy-covariance NEE and chamber-based component flux measurements.

Keywords: localized near field theory, inverse methods, ecosystem respiration, net ecosystem exchange, forest carbon budget

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Introduction

Estimating monthly to annual ecosystem respiration (Rₑ) is essential in light of the pre-eminence of respiration in affecting net ecosystem CO2 exchange (Baldocchi et al. 1997; Valentini et al. 2000). Long-term covariance measurements of net ecosystem exchange (NEE) are critical to estimating ecosystem carbon budgets, yet large errors arising from micrometeorological measurements of respiration under night-time conditions may corrupt NEE estimates (e.g. Wofsky et al. 1993; Baldocchi et al. 1996; Goulden et al. 1996; Moncrieff et al. 1996; Lindroth et al. 1998; Law et al. 1999a,b; Schmid et al. 2000; Valentini et al. 2000). Vertical subsidence, lack of

Correspondence: Chun-Ta Lai, tel. +801 581 8917, fax 801 581 4665, e-mail lai@biology.utah.edu

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steadiness in mean atmospheric conditions and intermittent turbulent transport contribute to increase the ‘de-coupling’ between night-time net ecosystem respiration and the CO₂ flux above the canopy measured by covariance methods. Instrument limitations, such as separation between sensors that are assumed to measure the same parcel of air, attenuation of fluctuation in the tube, volume averaging by anemometers, and finite sampling periods that may be too short to resolve all the eddy sizes, further compound the random and systematic errors in night-time flux measurements (Kaimal & Gaynor 1991; de Bruin et al. 1993; Kaimal & Finnigan 1994; Leuning & Judd 1996; Moncrieff et al. 1996; Massman 2000).

An alternative approach to estimating \( R_E \) is to utilize a functional relationship between CO₂ sources, \( S(t,z) \), or fluxes, \( F(t,z) \), and a relatively simple quantity to measure such as mean CO₂ concentration, \( C(t,z) \), within the canopy volume, where \( t \) is time and \( z \) is the height from the forest floor. This approach has the added benefit of permitting detailed partitioning of \( S(t,z) \) between different canopy levels and between the vegetation and the forest floor efflux \( F(t,0) \).

Early research efforts derive these functional relationships by assuming \( F(t,z) \) is related to \( dC/dz \) via an effective turbulent diffusivity (K-Theory). Over the past 30 years, theoretical developments and many laboratory and field experiments have demonstrated that scalar and momentum fluxes within many canopies do not obey K-Theory (e.g. Deardorff 1972, 1978; Corrsin 1974; Shaw 1977; Sreenivasan et al. 1982; Denmead & Bradley 1985; Finnigan 1985; Raupach 1988; Wilson 1989). To alleviate K-Theory limitations, other theoretical and practical methods were developed without resorting to a local eddy diffusivity approximation (e.g. Raupach 1988, 1989a,b; Katul & Albertson 1999; Katul et al. 2000a; Siqueira et al. 2000).

This study builds on these recent advances to develop a method capable of inferring \( F(t,0) \) and \( S(t,z) \) from measured \( C(t,z) \) for long-term estimation of night-time \( R_E \). To successfully reproduce \( F(t,z) \) from measured \( C(t,z) \), the following conditions must be satisfied: (i) accurate description of the flow statistics inside the canopy for stable atmospheric conditions, (ii) explicit accounting of storage fluxes within the canopy volume, and (iii) constraining the estimation of \( S(t,z) \) so as to minimize the effect of inherent measurement and sampling errors in \( C(t,z) \) on the computed \( S(t,z) \). The objective of this study is to develop a method, termed ‘Constrained Source Optimization’ or CSO, capable of reproducing night-time ecosystem fluxes and \( R_E \) components from measured \( C(t,z) \). The method is tested using 1 year of 30-minute measured mean CO₂ concentration profiles \( C(t,z) \) collected within a 17-year-old (in 1999) even-aged loblolly pine (Pinus taeda L.) stand at the Duke Forest. Rather than propose a replacement method to night-time covariance, chamber-scale measurements, or other established respiration measurement methods, we propose here a complementary method that can independently assess monthly to annual \( R_E \) estimates from routinely measured \( C(t,z) \). Such an assessment is essential because forest floor respiration is a major component of forest gross primary production (Baldocchi et al. 1997; Law et al. 1999b; Valentini et al. 2000).

Theory

In order to estimate \( R_E \), both the source \( S(t,z) \) in the canopy and the forest floor CO₂ flux \( F(t,0) \) are required because

\[
R_E = \int_0^h S(t,z) \, dz + F(t,0) \tag{1}
\]

where \( h \) is the mean canopy height. Measured \( C(t,z) \) can be related to \( S(t,z) \) or \( F(t,z) \) using the time and horizontally averaged one-dimensional continuity equation for a planar homogeneous flow,

\[
\frac{\partial C(t,z)}{\partial t} = -\frac{\partial F(t,z)}{\partial z} + S(t,z) \tag{2}
\]

Equation 2 has two unknowns, \( S(t,z) \) and \( F(t,z) \), and requires an additional prognostic equation to solve for \( R_E \). One approach to establish this requisite equation is to consider the interdependency between \( S(t,z) \) and \( C(t,z) \) via Lagrangian dispersion theory, which is considered next.

Lagrangian Dispersion Theory

The localized near-field theory (LNF) proposed by Raupach (1989a,b) can be used to describe a canonical relationship between \( S(t,z) \) and \( C(t,z) \). According to Raupach (1989a,b), \( C(t,z) \) can be decomposed into far-field \( (C_f) \) and near-field \( (C_n) \) concentrations, with the near-field concentration at \( z \) given by

\[
C_n(z) = \sum_{k=0}^{n} S(z_0) \left\{ k_n \left[ \frac{z - z_0}{\sigma(w(z_0))} \right] \right. \\
+ \left. k_n \left[ \frac{z + z_0}{\sigma(w(z_0)T(z_0))} \right] \Delta z_0 \right\} \tag{3}
\]

where \( \Delta z_0 \) is the vertical interval between source locations, \( \sigma_n(z) \) is the standard deviation of the vertical velocity at height \( z \), \( k_n(\zeta) = -0.3989 \ln[1 - e^{-K}] - 0.15623e^{-K} \) is a near-field kernel function, and \( \zeta \) is a dummy variable. This analytical approximation to the kernel and resulting errors (< 6%) are discussed in Raupach (1989a,b, 1998) and Gu (1998). Given a reference concentration

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(C) at \( z/h > 1 \), the diffusive, far-field concentration \( C_f \) can be expressed as:

\[
C_f(z) = C_r(z_R) - C_n(z_R) + \sum z_k F(z) \Delta z
\]

where \( z_R \) is a reference (or measurement) height, and

\[
K_f(z) = \sigma^2(z)/T_l(z)
\]

is the far-field eddy diffusivity in a steady, homogeneous turbulent flow, and \( T_l \) is the Lagrangian integral time scale. We emphasize here that while steady-state diffusion is assumed in equations 3 to 5, the mass conservation equation in equation 2 is non-steady.

Many laboratory and field experiments demonstrated that LNF is a significant improvement over gradient-diffusion theory (Raupach et al. 1992; Denmead & Raupach 1993; Denmead 1995; Katul et al. 1997a; Massman & Weil 1999; Leuning 2000; Leuning et al. 2000; Siqueira et al. 2000). Therefore, LNF was used to describe the scalar transport in the CSH. Other models equally suitable to CSO include the Eulerian closure approach of Katul & Albertson (1999) and Katul et al. (2000a), the Hybrid approach of Siqueira et al. (2000), or the recent revision to LNF by Warland & Thurtell (2000).

For LNF, given \( T_l \) and \( \sigma_{om} \) equations 3–5 provide the independent formulation necessary to relate \( C \) to \( S \) and \( F \), thereby 'closing' equation 2 and, in principle, permitting night-time \( R_E \) to be computed from measurements of \( C \). As discussed in Katul et al. (1997a, 2000a), the conventional inverse LNF approach is not sufficiently 'robust' to measurement errors because, in practice, the difference between the number of source layers and the number of concentration measurement layers is typically small. Naturally, such a small difference implies limited degrees of freedom (usually \( \leq 3 \)) in the estimation of \( S \) from \( C \). We describe next the CSO, which is a variant on Raupach’s (1989b) inversion to generate the desirable robustness in the inference of \( S \) from \( C \).

**Constrained Source Optimization (CSO) method**

As described by Raupach (1989b), it is possible to solve for the source \( S \) and the flux \( F \) from the mean CO\(_2\) concentration profile \( C \) by the so-called 'inverse' approach. In this method, the canopy is first divided into \( m' \) source layers, where \( m' \leq m \), and \( m \) is the number of levels where \( C(z) \) are measured \((j = 1, 2 \cdots m)\). The challenge is to determine the optimum source distribution \( S(z) \) \((j = 1, 2 \cdots m')\) that best recovers the measured \( C(z) \) using LNF and prescribed flow statistics (e.g. \( \sigma_{om}(z), T_l(z) \)). As discussed in Katul et al. (1997a) and Siqueira et al. (2000), this inverse procedure is very sensitive to the choice of \( m' \) and \( m \), as well as to the random measurement errors in \( C(z) \). With \( m - m' < 3 \) for most field studies, spurious sources and sinks may be produced from measurements and sampling errors in \( C \) (see Katul et al. 2000a; Siqueira et al. 2000). To minimize such sensitivity and increase the degrees of freedom in the estimation of \( S \) from \( C \), an alternative approach that utilizes known canopy physiological properties and attributes is proposed.

Total above-ground plant respiration \((= \int_0^h S(t, z)dz)\) can be decomposed into woody tissue and foliage respiration \((F_f)\). According to Law et al. (1999b), wood respiration accounted for about 6% of total \( R_E \) in a ponderosa pine ecosystem, while foliage respiration accounted for about 18%, despite the relatively low leaf area density in that forest. In our region, atmospheric demand for vapour is lower than that in the Oregon ponderosa pine ecosystem, thus requiring less woody water transport system per unit of leaf area (Oren et al. 1986). Therefore, we assumed the contribution of wood respiration would be even less important relative to the foliage respiration in comparison to the ponderosa pine ecosystem. Consequently, the contribution of woody components to above-ground ecosystem respiration was not explicitly separated from foliage in this study. In a nearby stand of similar characteristics, the contribution of woody surface area to plant area density was 18% (Pataki et al. 1998). This is an underestimate of the total woody surface area in the stand and hence the actual amount of respiring woody biomass. This was partially compensated for in the calculations of respiration by assuming that the entire plant area density was composed of foliage with its characteristically higher respiration rates.

With this simplification, a first order estimate of the vertical distribution of \( S \) can be specified \textit{a priori} as

\[
S(z) = a(z)R_d(z)
\]

where \( a(z) \) is the plant area density and \( R_d \) is the dark respiration per unit plant area modelled after Collatz et al. (1991) as

\[
R_d(z) = \alpha V_{cmax}(z)
\]

where \( \alpha \) is, in our approach, an unknown constant and \( V_{cmax}(z) \) is the maximum catalytic capacity of Rubisco per unit leaf area. The scaling between \( R_d \) and \( V_{cmax} \) assumes that respiration is mainly a function of enzyme turnover associated with the maintenance of photosynthetic enzymes, principally ribulose 1,5-bisphosphate carboxylase as represented by \( V_{cmax} \). Naturally, other models for \( R_d \) can be readily implemented in such a framework. Both \( V_{cmax} \) and \( \alpha \) are temperature-dependent (Campbell & Norman 1998). Typical \( \alpha \) values at 25°C are assumed between 0.011 and 0.015 (Farquhar et al. 1980; Collatz et al. 1991). Here, the temperature dependency of \( V_{cmax} \) can be calculated from
\[ V_{c, \text{max}}(T_a) = V_{c, \text{max}, 25} \exp[a_1(T_a - 25)] \frac{1}{1 + \exp[a_2(T_a - 29)]} \] (8)

where, \( V_{c, \text{max}, 25} \) is the maximum catalytic rate of Rubisco at 25°C, \( T_a \) is the air temperature (which approximates the plant tissue temperature during night time), and \( a_1 \) and \( a_2 \) are species-specific coefficients as discussed by Lai et al. (2000a), which were reported, respectively, as 0.051 and 0.208 for overstorey pine (0.4 < z/h < 1) and as 0.088 and 0.290 for the understory hardwood (0 < z/h < 0.4). Long-term averaged values of \( V_{c, \text{max}, 25} = 59\mu\text{mol m}^{-2}\text{s}^{-1} \) for pines and 30\mu\text{mol m}^{-2}\text{s}^{-1} \) for the hardwood species are used, as discussed by Ellsworth (1999), Delucia & Thomas (2000) and Naumburg & Ellsworth (2000).

We define the constrained source optimization (CSO) method as the combination of \( z \) and forest floor flux \( F(t,0) \) used in equations 2–5 to recover the measured \( C(z) \) in a least-square manner (e.g. optimized) every 30 min. When compared to Raupach’s (1989b) inverse method, this approach offers several advantages for night-time source calculations:

1. The proposed CSO modelling approach prescribed and constrained \( S(z) \) by our understanding of the vertical distribution of \( \alpha(z) \) and \( V_{c, \text{max}}(z) \), both measured quantities, but allowing determination of ecosystem respiratory fluxes \( F(\alpha(0),0) \) to reproduce the \( CO_2 \) concentration field for a wide range of stable atmospheric conditions. In essence, the number of unknowns was greatly reduced to \( m' = 2 \). The redundancy in the proposed optimization is greater (typically, \( m - 2 \) varies from 6 to 8) when compared to Raupach’s (1989b) method (\( m - m' \leq 3 \) for nearly all LNF field studies conducted thus far). Hence, with such redundancy, \( S \) and \( F(0,0) \) are no longer ‘hypersensitive’ to random measurement and sampling errors in \( C(z) \).

2. Because the number of parameters is small (only \( \alpha \) and \( F(0,0) \)), a global optimization method can be efficiently implemented. This global search can be readily constrained by the maximum range in \( \alpha \) and \( F(0,0) \) from literature values. For example, a global search with \( \alpha \) between 0.001 and 0.05, and \( F(0,0) \) between 0 and 30 \( \mu\text{mol m}^{-2}\text{s}^{-1} \), should yield reasonable sources and forest floor fluxes observed in majority of field experiments.

3. The method takes advantage of the measured plant area density \( a(z,t) \) dynamics when computing the canopy sources. Plant area density distribution and dynamics are not well utilized in the LNF approach of Raupach (1989a, b).

The velocity statistics and integral time scale

Ideally, the measured \( \sigma_w(z) \) should be used in equation 3 and 5. However, these measurements are rarely available on a routine basis on monthly to annual time scales. Higher-order Eulerian closure models provide an alternative to routine estimation of \( \sigma_w(z) \) provided that they can be validated via short intensive field campaigns. A key advantage to such a model framework is its ability to map variability in \( a(z,t) \) to variability in \( \sigma_w(z,t) \).

In order to model the \( \sigma_w \) within the canopy volume, we chose the second-order closure approach by Wilson & Shaw (1977) (hereafter referred to as WS77; see Appendix 1 for details). The ability of this model to reproduce measured \( \sigma_w \) inside forested canopies was found to be comparable to third-order closure schemes (Kutul & Albertson 1998) and other multiscale second order closure schemes (Kutul & Chang 1999). Furthermore, obtaining the numerical solution to the WS77 closure equations does not require complex numerical integration algorithms, at least when compared to the recently proposed closure model of Ayotte et al. (1999). Although the WS77 assumes neutral atmospheric conditions, the effects of atmospheric stability can be indirectly accounted for by revising the upper boundary conditions to evolve with \( h/L \), where \( L \) is the Obukhov length typically measured at a reference height above the canopy–atmosphere interface.

With regards to the integral time scale, Raupach (1989a, b) suggested that \( \frac{T_{1,u}}{h} \approx 0.3 \) as a constant within the canopy volume, where \( u^* \) is the friction velocity. While the 0.3 value appears reasonable for unstable and near-neutral atmospheric stability conditions (Kutul et al. 1997a), little evidence is available to support such a constant value for stable conditions. In this study, we assess the sensitivity of the \( R_b \) calculation to three different formulation of \( \frac{T_{1,u}}{h} \):

\[
\frac{T_{1,u}}{h} = \begin{pmatrix} 0.3 \\ \phi_1 \frac{h}{L} \\ \phi_2 \frac{h}{L} \end{pmatrix}
\] (9)

where, \( \phi_1 \) and \( \phi_2 \) are determined from detailed Eulerian vertical velocity integral time-scale measurements \( T_w \) and assuming \( T_w \approx \frac{1}{3} T_L \) (Tennekes & Lumley 1972). Estimates of \( T_w \) can be derived from the autocorrelation function \( \rho_{u^*}(t) \) of the vertical velocity fluctuations \( u^* \):

\[
T_w = \int_0^\infty \rho_{u^*}(t)dt, \quad \text{and} \quad \rho_{u^*}(t) = \frac{1}{\sigma_w^2} \left( u^*(t)u^*(t + \tau) \right)
\] (10)

where overbar represents time averaging (i.e. \( \overline{u^*} = 0 \)), and \( \tau \) is the time lag. The above integration is commonly terminated at the first zero crossing of \( \rho_{u^*}(t) \). If such a crossing is not well defined (as may be the case for some
very stable runs), the integration is terminated at the first \( \tau \) for which \( \rho_{0}(t) \) is no longer statistically different from zero at the 99% probability level. This probability limit on \( \rho_{0}(t) \) can be determined from Salas et al. (1984):

\[
\rho_{99\%} (t) = \frac{-1 \pm 2.326 \sqrt{N - \tau \delta t} - 1}{N - \tau \delta t} \tag{11}
\]

where \( N \) is the number of sample points, and \( \delta t \) is the sampling frequency (e.g. \( N = 18000 \) for \( \delta t = 10 \) Hz and a 30 min sampling duration). Hence, for consistency, the integration in equation (10) is terminated when \( \rho_{0}(t) \approx \rho_{99\%}(t) \) (rather than zero) for all runs.

**Study site and measurements**

**Study site**

The measurements are conducted at the Blackwood Division of Duke Forest in Orange County, near Durham, NC, USA (36°2’N, 79°8’W) as part of the AmeriFlux long-term CO₂ flux monitoring initiative (Kaiser 1998). The site is a uniformly planted loblolly pine (Pinus taeda L.) forest that extends 300–600 m in the east–west direction and 1000 m in the north–south direction. The mean canopy height was 14 m (± 0.5 m) at the end of 1998. The change of topographic slope is small (< 5%) so that the turbulent transport is not confounded by terrain (Kaimal & Finnigan 1994). All the measurements and parameterizations in this study were from the tower in ambient conditions.

**Eddy covariance flux measurements**

The CO₂ latent and sensible heat fluxes above the canopy were measured by a conventional covariance system comprising of a Licor-6262 CO₂/H₂O infrared gas analyser (Li-Cor Inc., Lincoln, NE, USA) and a Campbell Scientific sonic anemometer (CSAT3, Campbell Scientific Inc., Logan, UT, USA). The sonic anemometer was positioned just above the canopy (\( z = 15.5 \) m). The infrared gas analyser was housed in an enclosure on the tower 4.5 m below the inlet cup, which was co-located with the anemometer. The sampling flow rate for the gas analyser is 9 L min⁻¹, sufficient to maintain turbulent flows in the tubing. A krypton hygrometer (KH2O, Campbell Scientific) was positioned with the anemometer to assess the magnitude of the tube attenuation and time lag between vertical velocity and scalar concentration fluctuations as discussed in Katul et al. (1997a, 2000a).

The flux measurements were sampled at 10 Hz using a Campbell Scientific 21X data logger with all digitized signals transferred to a portable computer via an optically isolated RS232 interface for future processing.

Details about the processing and corrections to these high frequency measurements are described in Katul et al. (1997a,b).

**Mean CO₂ and water vapour concentration profiles within the canopy**

Another LiCor-6262 gas analyser was used to measure CO₂ and water vapour concentration profiles at 10 levels (\( z_i = 0.1, 0.5, 1, 1.5, 3.5, 5.5, 7.5, 9.5, 11.5, 13.5 \) and 15.5 m, i.e. \( m = 10 \)). This profiling system includes a multipoint gas sampling manifold to sample each level for 1 min (45 s sampling and 15 s purging) at the beginning, the middle, and the end of 30-minute sampling duration. In 1999, the experiment resulted in 4850 30-minutes night-time runs for which simultaneous CO₂ concentration and flux measurements were available between 2030 and 0430 EST.

**Other meteorological and biological variables**

In addition to the covariance flux measurements above the canopy, a HMP35C Ta/RH probe (Campbell Scientific) was also positioned at \( z = 15.5 \) m to measure mean air temperature (\( T_a \)) and mean relative humidity. A nonlinear thermistor (Campbell Scientific) was positioned at 15 cm below the forest floor, adjacent to the tower, to measure mean soil temperature (\( T_s \)). All the meteorological variables were sampled at 1 s and averaged every 30 min using a 21X Campbell Scientific data logger. A total of 45 leaf-level CO₂ assimilation (\( A_{\text{net}} \)) responses to CO₂ supply (\( A_{\text{net}} - C_\text{i} \) curves), conducted within a 14-month period between 1998 and 1999, were also used to estimate \( V_{\text{cma}} \), using the least-square regression procedure described by Wullschleger (1993) and Ellsworth (2000), where \( C_\text{i} \) is the leaf internal CO₂ mole fraction. Further details about the gas exchange measurements and sampling methodology can be found in Ellsworth (1999, 2000) and Katul et al. (2000b).

**Plant area density**

The vertical variation of the leaf area plus branches was measured by gap fraction techniques as described by Norman & Welles (1983). A pair of optical sensors with hemispherical lenses (LAI-2000, Li-Cor) was used for canopy light transmittance measurements from which gap fraction and plant area densities were calculated. The measurements were made at 1 m intervals from the top of the canopy to 1 m above the ground to produce the vertical profile in plant area index (PAI). PAI measurements were made routinely during the experimental periods, from the same tower used for measuring flow statistics, CO₂ concentration and covariance.
measurements. For time increments between these measurements, the PAI values were linearly interpolated. The vertical foliage distributions \( a(z) \) were generated using three canonical PAI profiles representing winter (December to April), summer (June to October) and transition periods (May and November).

**Soil respiration measurements**

Soil respiration \( F(t,0) \) was measured with a custom designed, chamber based system – Automated Carbon Eflux System (ACES) – based on principles used in Maier & Kress (2000). ACES is a multiport, dynamic gas sampling system that utilizes an open flow-through design to measure carbon dioxide fluxes from the forest floor. Each chamber was equipped with a pressure equilibration port, ensuring that the chamber pressure was held near ambient (Fang & Moncrieff 1996). Fifteen soil chambers (25 cm diameter, 10 cm height, 4909 cm\(^3\)) were measured sequentially using a single infrared gas analyser (EGM-3, PP Systems Inc., Haverhill, MA, USA) integrated into the system. Each chamber was sampled for 10 minutes to ensure steady state conditions, and a value of \( F(t,0) \) was recorded on the tenth minute of the cycle. Nine complete runs (135 measures) were recorded every day. When not being actively sampled, all chambers were continuously supplied with ambient air to prevent CO\(_2\) build-up. Data were collected continuously from April to May 2000. To lessen chamber induced impacts on the soil surface, chambers were moved twice a week between two fixed sample points.

**Velocity statistics measurements for testing WS77**

To estimate \( \phi_1 \) and \( \phi_2 \) in equation 9 and test the WS77 model performance for stable atmospheric conditions, an intensive field experiment was conducted. The three velocity components and virtual potential temperature inside the canopy were measured at eight levels \( (z = 1.3, 2.9, 4.8, 6.5, 8.5, 10.7, 12.1, 15.5 \text{ m}) \) from 23–30 August 1999 with a vertical array of eight CSAT3 sonic anemometers. Each sonic anemometer was anchored on a horizontal bar extending 1 m away from the walkway tower. The sampling frequency was 10 Hz for all sonic anemometers. All analogue signals were digitized by a Campbell Scientific CR-9000 data logger and transferred to a laptop computer for future processing. This experiment was performed at the same flux tower site used to measure plant area density, the long-term scalar fluxes and concentration profiles thereby permitting to assess the WS77 performance of reproducing \( \sigma_w(z) \) using the locally measured \( a(z) \). The velocity statistics were collected over a wide range of atmospheric stability conditions; however, here we focus only on stable atmospheric conditions.

**Results and discussion**

We assessed the performance of the CSO by (i) comparing the predicted night-time ecosystem fluxes with eddy covariance measurements collected at \( z/h = 1.1 \) for high \( u_\ast \) \( (> 0.15 \text{ m s}^{-1}) \), (ii) verifying whether the optimized \( \alpha \) is bounded by leaf-level measurements, and (iii) evaluating the similarity of forest floor respiration–soil temperature response function computed with the CSO to functions derived from chamber measurements. Finally, the seasonal dynamics of \( R_E \) and its components, as modelled by the CSO, are discussed in the context of the annual carbon budget and gross primary production (GPP) of this forested ecosystem. Before presenting the modelled \( R_E \) results, we consider the estimation of the flow statistics inside the canopy that are essential to drive the CSO method.

**Velocity statistics for nocturnal conditions**

To model night-time \( R_E \) using LNF, the velocity statistics \( \sigma_w \) and \( T_L \) are required. Figure 1 shows the measured \( T_L \) and \( \sigma_w/u_\ast \) at \( z/h = 1.1 \) for a wide range of stable atmospheric conditions. Here, \( T_L \) is derived from the Eulerian integral time scale (\( T_{ei} \)) described in equation 10. The apparent increase of \( \sigma_w/u_\ast \) from the value of 1.1 for large \( h/L \) is attributed to the fact that random errors tend to increase \( \sigma_w \) but decrease \( u_\ast \) over very stable conditions. Hence, we assumed that \( \sigma_w/u_\ast = 1.1 \) for the entire \( h/L \) range. This is consistent with the findings of Leuning (2000) that thermal stratification has little effect on \( \sigma_w \).

While the normalized \( \sigma_w \) remains nearly constant (c. 1.1), the \( T_L \) is greatly reduced with increasing stability (a reduction factor of 10 for two decades of \( h/L \)). Based on these field observations, \( T_L \) was made to vary with \( h/L \). This variation is also consistent with findings in a recent study by Leuning (2000) in which the LNF performance was significantly improved when \( T_L \) was corrected based on atmospheric stability. To quantify the stability effects on \( T_L \), we fitted an exponential curve \( f(h/L) \) to the measurements in Fig. 1. The fitted curve was used to compute \( T_L \) based on measured \( h/L \) at \( z/h = 1.1 \). The above analysis completes the description of the upper boundary condition for the canopy volume in order to generate the flow statistics for LNF.

The within-canopy vertical variation of \( \sigma_w \) and \( T_L \) are considered using the intensive velocity statistics measurements collected inside the canopy. Figure 2 shows the normalized ensemble \( \sigma_w \) and \( T_L \) profiles inside the canopy for nocturnal conditions collected in August 1999. Although WS77 does not explicitly account for density gradients, it can reproduce night-time \( \sigma_w \) reasonably well.
Fig. 1 Normalized \( T_L \) (top) and \( \sigma_w \) (bottom) measured (○) at \( z/h = 1.1 \) for different stable conditions. Solid line is the fitted curve used in the CSO calculation. Dashed lines are the values for near-neutral conditions within the canopy sublayer.

Finally, the case for which \( T_L \) varies with both atmospheric stability \( h/L \) and \( z/h \) as derived from \( T_w \) measurements is also shown. This case can be approximated by

\[
3. \quad \frac{T_L u_*}{h} = \phi_2 \left( \frac{h}{L}, \frac{z}{h} \right) = f_{g_2} - 0.155 \times (z_c - 1),
\]

where

\[
\begin{align*}
    z_c &= 0.4 & \text{if } z/h < 0.4 \\
    z_c &= z/h & \text{if } z/h \geq 0.4
\end{align*}
\]

Fig. 2 Comparison between ensemble averages of within canopy \( T_L \) (right) and \( \sigma_w \) (left) profiles measured in August 1999 and modelled based on WS77. For reference, the profiles suggested by Raupach (1989a,b) are also shown.

Fig. 3 Three plausible \( T_L \) profiles tested in the CSO approach. For the \( T_L u_*/h = f_g(\phi_1, \phi_2) \), we present as a sample \( T_L u_*/h = 0.044 \) at \( z/h = 1.1 \) derived from the long-term mean value in Fig. 2.

if the empirical constant in the characteristic length scale is varied from its neutral value to match best the measured mean wind field (see Appendix 1). The \( \sigma_w \) profile from WS77, obtained with the revised length scale constant, is used in the LNF calculation for the entire study period.

While the vertical variation in \( T_L \) cannot be directly measured, it may be inferred from measured \( T_w \) shown in Fig. 2. The fact that the measured \( T_w \) profile is not constant within the canopy volume motivated us to consider three possible scenarios for the space-time variation in \( T_L \), shown in Fig. 3, and described below in the following three cases.

1. \( \frac{T_L u_*}{h} = 0.3 \), constant throughout the canopy without any stability correction.
2. \( \frac{T_L u_*}{h} = f_0 \), where \( f_0 = \phi_1 \left( \frac{z}{h} \right) = \exp (-0.0695 \times 0^2 - 0.6454 \times -3.8516) \), \( x = \ln \left( \frac{z}{h} \right) \), and \( L \) is calculated from the sensible heat and friction velocity measurements collected at the canopy top. Here, \( T_L \) is also assumed constant within the canopy but varies with atmospheric stability at \( z/h = 1.1 \).
Again, we emphasize that this profile assumes $T_w \approx \frac{1}{3} T_L$ and is based on the measurements shown in Fig. 2. From the analysis in Appendix 2, we found that case 2 provides the most reasonable result, although the differences between cases 2 and 3 was at most 10%. For the remaining discussion the $T_L$ profile of case 2 was used.

Model optimization

Given the modelled velocity field and assuming some $\alpha$ and $F(t,0)$, the night-time CO$_2$ concentration profile can be computed using equations 3–5. The $\alpha$ and $F(t,0)$ are then allowed to vary until the root-mean square error (RMSE) between predicted and measured CO$_2$ concentrations at 9 levels is minimized.

The optimization was conducted for each 30 min run with

1. $\alpha > 0$, (i.e. stable atmospheric conditions)
2. measured $F(t,0) > 0$ (ecosystem respiration exceeds photosynthesis), and
3. RMSE < 10$\mu$ mol m$^{-1}$ (i.e. LNF reproduces well the measured concentration).

These three criteria resulted in 2418 30-minute evening runs between 2030 and 0430. The resulting optimized concentration profiles are shown in Fig. 4 after ensemble averaging every 2 months (to illustrate seasonal patterns).

As evidenced from Fig. 4, the C profiles computed with CSO from optimized $\alpha$ and $F(t,0)$ are in good agreements with the ensemble-averaged C measurements.

Model validation

To assess how well CSO reproduced $R_D$, we consider the following:

1. Comparisons between predicted and measured ensemble of $F(t,h)$ for $u_\tau > 0.15$ m s$^{-1}$. The latter $u_\tau$-condition is used to ensure that the measured $F(t,h)$ reliably represents net ecosystem exchange (NEE) (Falge et al. 2000).
2. A comparison between modelled forest-floor-temperature function and chamber-measured response functions.
3. Whether the optimized $\alpha$ value is bounded by the porometry chamber measurements of leaf $R_D$.
4. On longer time scales, whether the modelled ratio of forest floor to total ecosystem respiration is consistent with independently established estimates derived from biomass increment and other ecological measurements.

Comparisons between modelled and covariance measured $F(t,h)$

A number of previous studies have found that during night-time conditions, the NEE measurements by eddy covariance may not be reliable (e.g. Wofsy et al. 1993; Hollinger et al. 1994; Black et al. 1996; Goulden et al. 1996; Baldocchi et al. 1997; Lavigne et al. 1997; Law et al. 1999a). For example, when eddy covariance measurements were compared to the chamber-estimated soil surface efflux in a ponderosa pine forest in Oregon, the underestimation was as large as 50% (Law et al. 1999b). On the other hand, other field experiments did not find such discrepancies (Grace et al. 1996). Uncertainty in these comparisons may be complicated by the storage flux ($F_{st}$), which is given by

$$F_{st} = \int_{0}^{h} \frac{\partial C}{\partial t} dz$$

(12)

Greco & Baldocchi (1996) reported that $F_{st}$ averages to near zero on a daily basis, although it may be significant.
in the evenings. Previous measurements at this site also showed that \( F_{\text{st}} \) averages to zero on a daily basis but is significant during sunrise and sunset periods (Lai et al. 2000b). For this study, \( F_{\text{st}} \) was calculated using measured mean CO\(_2\) concentration every 30 min for the selected night-time runs (from 2030 to 0430) and was included for reference in the comparison between measured and modelled ecosystem respiration.

Upon integrating equation 2 with respect to \( z \), we obtain

\[
\int_{0}^{h} \frac{\partial C(t, z)}{\partial t} \, dz + F(t, h) = F(t, 0) + \int_{0}^{h} S(t, z) \, dz, \tag{13}
\]

in equation 13, the left-hand side represents the measured ecosystem respiration and the right-hand side represents the so-called 'biotic' fluxes. To illustrate the seasonal variation in these components, we report the ensemble averages of measured \( F_{\text{st}} + F(t, h) \) and CSO-modelled \( F(t, 0) + \int_{0}^{h} S(t, z) \, dz \) every 2 months for evening runs between 2030 and 0430 in Fig. 5. For \( u > 0.15 \text{ m s}^{-1} \) (a threshold used by several investigators including Goulden et al. 1996 and Clark et al. 1999), our model calculation agrees closely with measured \( F_{\text{st}} + F(t, h) \) for the winter periods. However, modelled \( F(t, 0) + \int_{0}^{h} S(t, z) \, dz \) tends to be greater than measured \( F_{\text{st}} + F(t, h) \) by up to 30% in the summer. This discrepancy may be an artifact of the modelled time scale \( (T_s) \), or an indication of missing fluxes which are not well captured by either eddy covariance system or the profiling system used to estimate the storage flux. Generally \( F_{\text{st}} \) are small except for summer months, when mean CO\(_2\) concentration near the forest floor \( (z < 10 \text{ cm}) \) is large and dynamic. We emphasize that both measured \( F(t, h) \) and \( F_{\text{st}} \) suffer from random and sampling errors.

**Optimized forest-floor respiration**

In order to compare the CSO model results with other studies and chamber measurements available at this site, the CSO-optimized forest floor respiration are presented as a function of soil temperature (e.g. Raich & Schlesinger 1992; Wofsy et al. 1993; Hollinger et al. 1994; Lloyd & Taylor 1994; Fan et al. 1995; Lindroth et al. 1998; Law et al. 1999a,b). Regressing the CSO modelled \( F(t, 0) \) with soil temperature \( (T_s) \) measured at 15 cm below the ground surface resulted in

\[
F(t, 0) = 1.198 \exp (0.054 T_s) \tag{14}
\]

This regression was performed with bin-averaged respiration values for 5°C increments of \( T_s \) (Fig. 6a). Figure 6(b) shows the comparison between this function and functional relationship derived from the ACES chamber measurements, along with functions reported in other studies. We note that nearly 70% of our measured \( T_s \) is between 5°C and 15°C, where the discrepancy between modelled and measured \( F(t, 0) - T_s \) is no greater than 7% (Fig. 6b). Such discrepancy can be attributed to a combination of model underestimation and statistical uncertainty in aggregating chamber measurements (e.g. Lavigne et al. 1997).

We estimated the \( Q_{10} \) value of the CSO-modelled \( F(t, 0) - T_s \) relationship in two ways. The first utilizes all the 30-minute runs and the second is based on bin-averages of modelled respiration and soil temperature every 5°C (Fig. 6a). The \( Q_{10} \) values derived from the CSO and the ACES chamber measurements are presented in Table 1. The agreement between the chamber-measured and CSO-modelled \( Q_{10} \) is within 6%. The result from the bin-averaged approach is better behaved and is used throughout. Both \( Q_{10} \) values derived from ACES chamber measurements and the CSO are also well within the range reported in other studies (Raich & Schlesinger 1992; Lavigne et al. 1997). However, these \( Q_{10} \) estimates are lower than those calculated using data from Andrews &

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Table 1 Estimates of \(Q_{10}\) and \(R_{10}\) from different experiments at the pine forest stand. \(R_{10}\) is the forest floor respiration at 10°C. Andrews & Schlesinger’s (2001) \(Q_{10}\) and \(R_{10}\) are derived from SRC-1 chamber measurements (PP Systems) collected at three plots (FOUR chambers per plot) within the forest, once a month (at noon) from 1996 to 1999. The ACES estimates of \(Q_{10}\) and \(R_{10}\) are from 30-minute forest floor flux and soil temperature \((T_s)\) measurements conducted in April–May (60 days) at one plot (14 chambers per plot). The range of values from Andrews & Schlesinger (2001) are due to spatial variability in \(Q_{10}\) and \(R_{10}\) among plots.

<table>
<thead>
<tr>
<th>Reference/approach</th>
<th>(Q_{10}/R_{10}) using (T_s \geq 83) cm</th>
<th>(Q_{10}/R_{10}) using (T_s = 15) cm</th>
</tr>
</thead>
<tbody>
<tr>
<td>Andrews &amp; Schlesinger (2001)</td>
<td>2.3–2.7/2.5–2.6</td>
<td>2.3–2.9/2.5–2.6</td>
</tr>
<tr>
<td>Automated Chambers (ACES)</td>
<td>****</td>
<td>1.8/2.1 (April-May 2000, 14 chambers)</td>
</tr>
<tr>
<td>CSO Model</td>
<td>****</td>
<td>1.7/2.0 (using 5°C bin averages)</td>
</tr>
<tr>
<td></td>
<td></td>
<td>1.5/1.9 (using all 30 min runs)</td>
</tr>
</tbody>
</table>

Fig. 6 (a) The 30-minute CSO modelled \(f(t,0)\) as a function of measured \(T_s\) (dots) at 15 cm below the soil surface. Bin-averaging of \(f(t,0)\) along \(T_s\) in 5°C increments is also shown (with one standard deviation). The dashed and solid lines represent the best-fit exponential functions to the raw 30-minute \(f(t,0)\) and to the bin-averaged \(f(t,0)\), respectively. (b) Comparison between the ACES measured and CSO modelled \(f(t,0)\) – \(T_s\) curves (bin-averaged). The respiration functions of Law et al. (1999a,b) and Lloyd & Taylor (1994) are also presented for reference. For the Lloyd & Taylor (1994) equation, the fitted regression coefficient \((T_0 = 209.58° K)\) to the bin-averaged CSO modelled \(f(t,0)\) was used.

Schlesinger (2001) for the same stand. The data of Andrews & Schlesinger (2001) resulted in \(Q_{10} = 2.9\) in the vicinity of the flux tower and a mean of 2.5 (when soil temperature measurements at 15 cm are used) for the stand in general (see Table 1). These estimates, higher by more than 30% of all the estimates in Fig. 6, may reflect biases in the measurement system (SRC-1, PP Systems, Haverhill, MA, USA) used by Andrews & Schlesinger (2001). Such discrepancies have been observed in direct field comparisons of the SRC-1, Li-Cor 6400-09 (Li-Cor, Inc.) and ACES (J. Butnor, unpublished data).

**Optimized \(\alpha\) and plant respiration**

The optimized 30-minute \(\alpha\) values were averaged over two-month period to obtain mean bi-monthly \(\alpha\) values for the entire year. A constraint on \(\alpha\) could be derived from daytime \(A_{net} - C_i\) curves collected around noon for the upper foliage (\(z = 12\) m). Upon fitting these curves to the Farquhar et al. (1980) model, the maximum possible \(\alpha\) can be determined after \(R_d\) and \(V_{c^{max}}\) are computed. Extrapolating the \(A_{net} - C_i\) curves to determine \(R_d\) from a small magnitude of \(A_{net}\) can, in itself, introduce systematic overestimation in \(R_d\) because of (i) the linear form of the extrapolating function and (ii) small and unavoidable leaks within gas analysers tend to increase the value of \(R_d\) (Villar et al. 1995; Amthor 2000). Based on the sunlit leaf-level \(A_{net} - C_i\) measurements and a linear extrapolation for \(R_d\), we found as expected that the mean \(\alpha = 0.025\) is larger than the CSO-optimized \(\alpha (= 0.012)\) by about 50%, for the same range in temperature in which the \(A_{net} - C_i\) curves were collected. We explored a quadratic extrapolation with the following constraints at the CO₂ compensation point (\(F\)).

\[
C_i = F_s \left\{ \frac{A_{net} - 0}{\frac{dA_{net}}{dC_i}} = V'_{c^{max}} \right\},
\]

and at \(C_i = 0\), \(dA_{net} = 0\), where \(V'_{c^{max}}\) is \(V_{c^{max}}\) computed with all kinetic corrections. The fundamental difference between linear and quadratic extrapolations is the slope condition at \(C_i = 0\), which assumes \(R_d\) independent of \(C_i\) for small \(C_i\) in the quadratic form. With this quadratic extrapolation, we calculated \(R_d\) that is 50% smaller than the linearly extrapolated \(R_d\) (Fig. 7) thereby reducing \(\alpha\) from 0.025 to 0.0125. The latter \(\alpha\) is in excellent agreement with the CSO-computed \(\alpha\), and is in accordance with reported physiological repression of respiration by leaf carbohydrates (Azcon-Bieto 1983; Villar et al. 1995). Both the quadratic interpolation for \(R_d\) and the CSO calculation produced estimates of \(\alpha\) within the range of reported values of 0.011–0.015 (e.g. Farquhar et al. 1980; Collatz et al. 1991).
Fig. 7 Estimation of dark respiration ($R_d$) by extrapolating daytime $A_{sat} - C_i$. The solid line is linear extrapolation leading to $\alpha \approx 0.025$. The dotted line is a quadratic extrapolation leading to $\alpha \approx 0.0125$ that is consistent with the CSO-modelled $\alpha$. The quadratic extrapolation includes repression of respiration by carbohydrates at very low $C_i$.

Seasonal variation of modelled plant and forest floor respiration

To summarize, we show the relative contribution of CSO-modelled plant and forest floor respiration to the total ecosystem respiration in Fig. 8a–d along with time-matching primary driving variables for different times of the year. The modelled $F(t,0)$ increased with mean soil temperature (Fig. 8c), reaching a near maximum rate of $3.8 \mu$mol m$^{-2}$ s$^{-1}$ in mid-summer (see Fig. 8a). The modelled plant respiration has a similar pattern, and also peaked in mid-summer with a maximum rate of $3.1 \mu$mol m$^{-2}$ s$^{-1}$ as a result of the high leaf area density (see Fig. 8d) and high plant tissue temperature. Figure 8b shows the ratio of modelled $F(t,0)/R_E$ in 1999 for different times of the year. This ratio varied from 0.82 to 0.56 from winter to summer, suggesting that the contribution of plant respiration to total ecosystem respiration is much less in the winter when both air temperatures and the leaf area density are low. These model results suggest that estimates of $F(t,0)/R_E$ derived from short field campaigns (even up to 2 months) should not be extrapolated to annual ratios.

Implications for annual ecosystem carbon balance

Based on the CSO-modelled $F(t,0) - T_s$ relationship, mean $\alpha$ values and measured daytime covariance fluxes, the major components of the annual carbon balance of the pine forest ecosystem can be quantified. To estimate daytime net CO$_2$ exchange, we regressed measured CO$_2$ fluxes with photosynthetically active photon flux density (PPFD) measurements using a nonlinear $F(t,h) - PPFD$ response curve (Landsberg 1986; Ruimy et al. 1995; Clark et al. 1999; Law et al. 1999a), given by:

$$F(t,h) = \frac{\alpha_p PPFD_{sat}}{\alpha_p PPFD + F_{sat}} - R_0$$

For this simple analysis, it is assumed that forest carbon assimilation is principally driven by light only with no other limitations (temperature, vapour pressure deficit, drought stress, etc.) explicitly considered. We employ this analysis here only for the purposes of comparing net C assimilation with other forest sites because light interception and radiation use are major determinants of forest productivity (Jarvis & Leverenz 1982; Ruimy et al. 1995). The estimated mean apparent quantum yield ($\alpha_p$), net CO$_2$ flux at light saturation ($F_{sat}$), and the mean net CO$_2$ flux at $PPFD = 0$ ($R_0$) are all close to the values reported by Clark et al. (1999) for a 24-year-old Florida slash pine forest of a similar leaf area index and height (see Table 2). The $F(t,h) - PPFD$ curve is shown in Fig. 9, and the regressed coefficients are compared in Table 2 with the Florida slash pine stand. The $F(t,h) - PPFD$...
equation was combined with continuous PPFD and other meteorological measurements to yield a net carbon gain of 1342 g C m\(^{-2}\) year\(^{-1}\), which is within 10% of the mean value reported by Clark et al. (1999).

Using equation 14 and measured soil temperature, \(F(0,0)\) was estimated to be 989 g C m\(^{-2}\) year\(^{-1}\), well within the range of temperate coniferous forest reported by Raich & Schlesinger (1992). DeLucia et al.'s (1999) study at this site reported a soil efflux as 1066 ± 46 and 928 ± 9 g C m\(^{-2}\) year\(^{-1}\) in 1997 and 1998, respectively, consistent with the CSO estimates. Based on the chamber-measured respiration-temperature curve of Fig. 6 (parameters in Table 1), \(F(0,0)\) = 1063 g C m\(^{-2}\) year\(^{-1}\), only 8% higher than the CSO modelled \(F(0,0)\).

Combining modelled \(\alpha\), measured plant area density and temperature-adjusted \(V_{\text{cmax}}\), \(F_j\) was estimated to be 214 g C m\(^{-2}\) year\(^{-1}\), slightly greater than the combined foliage and woody respiration reported by Law et al. (1999b) for a stand with a lower leaf area density than the loblolly pine ecosystem. The annual NEE of the loblolly pine forest, estimated as the difference between

![Fig. 9](image)

The whole-ecosystem light response (i.e. \(F(t, h)\) – PPFD relationship) derived from the eddy covariance measured \(F(t, h)\) at \(z/h = 1.1\) (O with one standard deviation) for the Pinus taeda ecosystem. For reference, the \(F(t, h)\) – PPFD curve for Pinus elliottii reported in Clark et al. (1999) is also shown (dot-dashed line).

**Table 2:** Comparisons of study site characteristics, \(F(t, h)\) and coefficients in \(F(t, h)\) – PPFD curve between two south-east pine ecosystems. The data for the North Carolina site are for 1999 (also see DeLucia et al. 1999 and Ellsworth 2000).

<table>
<thead>
<tr>
<th>Study site</th>
<th>North Carolina Lobolly pine forest</th>
<th>Florida Slash pine forest</th>
</tr>
</thead>
<tbody>
<tr>
<td>Location</td>
<td>36°2' N, 79°8' W</td>
<td>29°44' N, 82°9' W</td>
</tr>
<tr>
<td>Vegetation type</td>
<td>17-year-old Pinus taeda L.</td>
<td>24-year-old (in 1996) Pinus elliottii var. elliottii</td>
</tr>
<tr>
<td>Mean canopy height (m)</td>
<td>14 ± 0.5 (in 1998)</td>
<td>19.2 ± 1.0 (in 1996)</td>
</tr>
<tr>
<td>Maximum leaf area index</td>
<td>5.0</td>
<td>6.5</td>
</tr>
<tr>
<td>Minimum leaf area index</td>
<td>2.4</td>
<td>3.7</td>
</tr>
<tr>
<td>Site index at 25 years old</td>
<td>16 m</td>
<td>15 m</td>
</tr>
<tr>
<td>Long-term mean air temperature (°C)</td>
<td>7.6</td>
<td>14</td>
</tr>
<tr>
<td>Winter</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Summer</td>
<td>27.2</td>
<td>27</td>
</tr>
<tr>
<td>Long-term mean annual rainfall (mm)</td>
<td>1154</td>
<td>1332</td>
</tr>
<tr>
<td>(\Phi)</td>
<td>0.029</td>
<td>0.044</td>
</tr>
<tr>
<td>(F_{\text{sat}}) (µmol m(^{-2}) s(^{-1}))</td>
<td>31.2</td>
<td>26.5</td>
</tr>
<tr>
<td>(K_b) (µmol m(^{-2}) s(^{-1}))</td>
<td>3.9</td>
<td>4.6</td>
</tr>
<tr>
<td>Ecosystem light compensation point (µmol m(^{-2}) s(^{-1}))</td>
<td>156</td>
<td>125</td>
</tr>
<tr>
<td>Total daytime net (F(t, h)) exchange (g C m(^{-2}) year(^{-1}))</td>
<td>1342</td>
<td>1529 (1996)</td>
</tr>
<tr>
<td>Total plant respiration (g C m(^{-2}) year(^{-1}))</td>
<td>214</td>
<td>******</td>
</tr>
<tr>
<td>Total forest floor respiration (g C m(^{-2}) year(^{-1}))</td>
<td>989</td>
<td>1400 * (1995–96)</td>
</tr>
<tr>
<td>Total night time (F(t, h)) release (g C m(^{-2}) year(^{-1}))</td>
<td>737</td>
<td>789 (1996)</td>
</tr>
<tr>
<td>NEE (g C m(^{-2}) year(^{-1}))</td>
<td>1342 – 737 = 605</td>
<td>740 (1996)</td>
</tr>
<tr>
<td></td>
<td></td>
<td>608 (1997)</td>
</tr>
<tr>
<td>GPP (g C m(^{-2}) year(^{-1}))</td>
<td>214 + 989 + 605 = 1808</td>
<td>&gt; 2008</td>
</tr>
</tbody>
</table>

* Estimated by Moncrieff & Fang (1999) from chamber measurements and model simulations

daytime CO₂ gains and night time CO₂ release, was 605 g C m⁻² year⁻¹, also similar to that of the Florida slash pine forest. More importantly, this estimate (605 g C m⁻² year⁻¹) is also very close to the NEE estimated by the eddy covariance measurements corrected for mean u- with no gap-filling (630 g C m⁻² year⁻¹).

With these estimates of Rₑ and NEE, the estimated GPP = Rₑ + NEE is 1800 g C m⁻² year⁻¹ for 1999. For the purpose of this study, we define GPP to be carbon assimilation by photosynthesis ignoring photospiration (Schulze et al. 2000). To contrast this estimate with other independent GPP estimates, we consider the measured net primary production (NPP) along with the proposed NPP/GPP of ≈ 0.47 from Waring et al. (1998). Based on this ratio and an NPP of 800 g C m⁻² year⁻¹ reported by DeLucia et al. (1999) for this pine forest, GPP ≈ 1702 g C m⁻² year⁻¹, within 6% of the GPP determined from measured NEE and CSO-modelled Rₑ (see Table 3).

Another independent check on the forest GPP can be derived from the ratio of autotrophic respiration (Rₐ) to net photosynthesis (Aₐ = GPP). Using isotopic carbon labelling, Andrews et al. (1999) estimated the root contribution to be 55% of F(t,0). Combined with our estimate of F(t,0) ≈ 989 g C m⁻² year⁻¹, this yields a contribution by root respiration (Rₑ) of 544 g C m⁻² year⁻¹. When combined with the above-ground respiration (≈Fₑ = 214 g C m⁻² year⁻¹), Rₑ = Fₑ + Rₑ ≈ 758 g C m⁻² year⁻¹. Based on our estimate of GPP, Rₑ/GPP ≈ 0.42, which is well within the range of ratios reported by Ryan et al. (1994) for pine forests (0.33–0.63, with a mean = 0.49; see Table 3). Additionally, we consider the heterotrophic respiration RₐH estimated from 45% of F(t,0) as in Andrews et al. (1999) and the independently measured root turnover (TOₑ) and litter turnover (TOₑ) reported by DeLucia et al. (1999). Given the measurement uncertainty in these terms, the agreement between estimated RₐH (≈445 g C m⁻² year⁻¹) and measured TOₑ + TOₑ (≈400 g C m⁻² year⁻¹) is remarkably good (< 10%) as shown by the summary results in Table 3.

As a final test, the net photosynthesis Aₐ can be estimated from a simple Fick’s law formulation

\[
Aₐ = gₑCₑ \left[ \frac{1}{Cₑ} \right]
\]

where, gₑ is the bulk canopy conductance, directly estimated from sapflux measurements as described in Ewers et al. (1999) and Oren et al. (1998, 1999), Cₑ and Cₑ are the internal and atmospheric CO₂ concentrations, respectively. Using the approach of Lai et al. (2000a), the vertically averaged Cₑ/Cₑ equals 0.70 for the entire stand. The latter estimate is well within the range of the isotopic carbon analysis reported by Ellsworth (2000) and Katul et al. (2000b) for which Cₑ/Cₑ = 0.66 for sunlit and 0.76 for shaded foliage, respectively. The mean annual sapflux estimated gₑ of 70 mmol/m² s⁻¹ and Cₑ = 370 μmol mol⁻¹, leading to an estimate of Aₐ (=GPP, as per Schulze et al. 2000) of 1825 g C m⁻² year⁻¹ for 1999, which is in excellent agreement with the CSO model calculations.

Having demonstrated the consistency between our NEE measurements and modelled F(t,0) with a variety of experiments and literature data, we note that our estimated GPP (< 1800 g C m⁻² year⁻¹) is much larger (c. 25%) than the modelled GPP (c. 1200–1300 g C m⁻² year⁻¹) reported in Luo et al. (2001) for this pine ecosystem. In Luo et al. (2001), a multilevel photosynthesis model (MAESTRA) with an empirical parameterization of stomatal resistance, calibrated with porometry measurements collected at 12 m height, was used to estimate GPP. According to Luo et al. (2001), the MAESTRA model calculations reproduced measured NEE by the covariance system. Given that the MAESTRA model

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**Table 3** Comparison between ecosystem respiration components as modelled by CSO and estimated from published experiments. The units for respiration and GPP are in g C m⁻² year⁻¹

<table>
<thead>
<tr>
<th>Ecosystem Component</th>
<th>CSO Model</th>
<th>Literature Estimates</th>
</tr>
</thead>
<tbody>
<tr>
<td>GPP</td>
<td>1800</td>
<td>1700 derived from measured NPP = 800 by DeLucia et al. (1999) and estimated NPP/GPP = 0.47 by Waring et al. (1998).</td>
</tr>
<tr>
<td>Forest Floor Respiration</td>
<td>989</td>
<td>928 from DeLucia et al. (1999)</td>
</tr>
<tr>
<td>Total Root Respiration (Rₑ)</td>
<td>544</td>
<td>= 0.55 × 989, the 55% is from Andrews et al. (1999)</td>
</tr>
<tr>
<td>Autotrophic Respiration (Rₐ)</td>
<td>758</td>
<td>= 544 + 214, the 214 is from CSO modelled plant respiration</td>
</tr>
<tr>
<td>Rₐ/GPP</td>
<td>0.42</td>
<td>Rₑ/GPP = 0.33–0.63 reported by Ryan et al. (1994) for pine forests (mean = 0.49)</td>
</tr>
<tr>
<td>Heterotrophic Respiration (RₐH)</td>
<td>445</td>
<td>= 0.45 × 989; the 45% is from Andrews et al. (1999) Measured litter turnover = 300 (DeLucia et al. 1999) Measured root turnover = 100 (DeLucia et al. 1999) Hence, estimated Rₑ = 400 (for 1998 measurements)</td>
</tr>
</tbody>
</table>

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reproduced measured NEE (c. 630 g C m\(^{-2}\) year\(^{-1}\)), the model calculations must have underestimated \(R_E\). For instance, combining the measured NEE with measured forest floor respiration from DeLucia et al. (1999) (c. 928 g C m\(^{-2}\) year\(^{-1}\) in 1998) results in a GPP of at least 1558 g C m\(^{-2}\) year\(^{-1}\) without accounting for plant respiration (c. 200 g C m\(^{-2}\) year\(^{-1}\) based on the CSO model calculations). Furthermore, a GPP of 1200 g C m\(^{-2}\) year\(^{-1}\) and a measured NPP of 800 g C m\(^{-2}\) year\(^{-1}\) (DeLucia et al. 1999) produce a ratio of NPP/GPP = 0.66, 40% higher than the ratio reported by Waring et al. (1998), in contrast to good agreement with the results based on the CSO model. Thus the CSO model combined with simple photosynthesis formulations as well as the mass-balance of this forest suggest GPP values between 1800 and 2000 g C m\(^{-2}\) year\(^{-1}\).

Conclusions

This study proposed a method to estimate components of ecosystem respiration from readily available CO\(_2\) concentration profile measurements. Applying this method to a pine forest in North Carolina, we demonstrated the following:

1. For nocturnal conditions and close to the canopy-atmosphere interface, \(\frac{L_{w}}{h}\) was < 0.3 and sensitive to variations in atmospheric stability \(h/L\). However, \(\frac{L_{w}}{h}\) did not vary appreciably within the canopy (< 20%), at least when assuming \(T_L \approx T_w\).

2. Unlike \(\frac{L_{w}}{h}\), \(\sigma_{u}/u^*\) remained nearly a constant (= 1.1) for a wide range of \(h/L\) evaluated at the canopy top. This constant ratio of \(\sigma_{u}/u^*\) is consistent with a recent study by Leuning (2000) and perhaps suggests that modelling \(\sigma_{u}/u^*\) for nocturnal conditions may be simpler than unstable conditions.

3. The second-order closure model of Wilson & Shaw (1977) successfully reproduced the measured vertical variation of \(\sigma_{u}/u^*\) for a wide range of \(h/L\) conditions.

4. The combination of LNF and the proposed optimization scheme, termed CSO, reproduced well ecosystem respiration measured with the covariance system under conditions of large \(u^*\). The modelled forest floor respiration \(F(t,0)\) by the CSO approach was in close agreement with the chamber measurements when the night-time CO\(_2\) fluxes are described as a function of soil temperature.

5. Using ecosystem respiration modelled based on the CSO and daytime net CO\(_2\) exchange measured with eddy covariance, we estimated GPP = 1800 g C m\(^{-2}\) year\(^{-1}\) and NEE = 605 g C m\(^{-2}\) year\(^{-1}\) for this temperate lobolly pine forest. The ratio of measured NPP to modelled GPP (0.44) is close to the general ratio (0.47) by Waring et al. (1998). Furthermore, the modelled plant and forest floor respiration are consistent with measured litter and root turnover rates reported by DeLucia et al. (1999), and with isotopic analysis of root : microbial respiration reported in Andrews et al. (1999).

The broader implications of this study, when combined with the recent NEE measurements reported in Clark et al. (1999) from a Florida slash pine forest ecosystem, is that young (15- to 25-year-old) south-eastern pine forests appear to be among the largest atmospheric terrestrial carbon sinks when compared to other flux sites with published NEE data in Europe and North America (e.g. Wofsy et al. 1993; Hollinger et al. 1994; Greco & Baldocchi 1996; Valentini et al. 2000). These two pine forests along with the Picea sitchensis conifer plantation in Valentini et al. (2000) exhibit GPP values from 1800 to 2000 g C m\(^{-2}\) year\(^{-1}\). Warm temperate conifers at an age close to crown closure are likely to be among the most productive forest ecosystems because their leaf area index is close to maximum and the growing season is very long (Ellsworth 2000), assuring an efficient absorption of photosynthetically active radiation. In addition, according to Waring & Schlesinger (1985), the amount of non-photosynthetic biomass is not very large at this stage of stand developments, assuring limited carbon consumption in respiration.

Acknowledgements

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References


Appendix 1: Second Order Closure Model of Wilson & Shaw (1977)

The following is summarized from Katul & Albertson (1998, 1999) and Katul et al. (2000a); for completeness, the final mean momentum and Reynolds stress equations are repeated. In order to model the vertical variation in $\sigma_n$, the second order closure model of Wilson & Shaw (1977) is used and is briefly described next. Upon time and horizontally averaging the mean momentum and Reynolds stresses equations, and after performing the simplifications of (i) steady-state adiabatic flow, (ii) assuming that the form-drag by the canopy can be modelled as a drag force, (iii) neglecting the viscous drag relative to the form drag, (iv) closing all the triple-velocity products by a gradient-diffusion approximation (Donaldson & Du 1973; Mellor 1973; Mellor & Yamada 1974; Shaw 1977; Wilson & Shaw 1977; Wilson 1988; Wilson & Anderson 1990; Canuto et al. 1994; Abella & McFarlane 1997), (v) modelling the pressure-velocity gradients based on return-to-isotropy principles, (vi) assuming the viscous dissipation is isotropic and dependent on the local turbulence intensity (Mellor 1973), and (vii) assuming horizontal homogeneity, the second-order closure model of Wilson & Shaw (1977) reduces to:

\[
\begin{align*}
\text{Mean Momentum:} & \quad 0 = -\frac{du}{dz} - C_{do}(z) U^2, \\
\text{shear stress:} & \quad 0 = -2\nu U \frac{dU}{dz} + 2 \int (q_1 \frac{dU}{dz} \frac{dU}{dz}) + \frac{qU \frac{dU}{dz}}{3 \lambda_1} + C_{do} \frac{dU}{dz} U, \\
\text{Lagrangian Variance:} & \quad 0 = -2\nu U \frac{dU}{dz} + 2 \int (q_1 \frac{dU}{dz} \frac{dU}{dz}) + \frac{qU \frac{dU}{dz}}{3 \lambda_1} + C_{do}(z) U^3 \frac{dU}{dz}, \\
\text{Lateral Variance:} & \quad 0 = -\frac{dU}{dz} \left[q_2 \frac{dU}{dz} \frac{dU}{dz} \right] + \frac{qU \frac{dU}{dz}}{3 \lambda_2} + C_{do}(z) U^3 \frac{dU}{dz}, \\
\text{Vertical Variance:} & \quad 0 = -\frac{dU}{dz} \left[q_3 \frac{dU}{dz} \frac{dU}{dz} \right] + \frac{qU \frac{dU}{dz}}{3 \lambda_3} + C_{do}(z) U^3 \frac{dU}{dz}, \\
\end{align*}
\]

where $u$, $u_1 = u$, $u_2 = v$, $u_3 = w$ are the instantaneous velocity components along $x$, $y$, $z$ directions, $\lambda_1$, $\lambda_2$, and $\lambda_3$ are characteristic length scales for the triple-velocity correlation, the pressure-velocity gradient correlation, and viscous dissipation, respectively (Katul & Chang 1999), $\kappa (= 0.4)$ is Von Karman’s constant, $C_d$ is the foliage drag coefficient, $\gamma$ is an empirical constant, and $C_1$, $C_2$, $C_3$, and $C_n$ are closure constants that can be determined such that the flow conditions well above the canopy reproduce established atmospheric surface layer similarity relations. With estimates of these five constants ($C_1$, $C_2$, $C_3$, $C_n$, and $\gamma$), the five ordinary differential equations listed above can be solved for the five flow variables $U$, $U^2$, $Uv^2$, $Uw^2$ if appropriate boundary conditions are specified. The flow boundary conditions are assumed to approach the theoretical values of neutral atmospheric surface layer similarity (Panofsky & Dutton 1984) well above the canopy ($z/f > 2$), and zero gradients for all second moments, as well as zero mean velocity at the forest floor. Table 4 summarizes the closure constants used in WS77.

<table>
<thead>
<tr>
<th>Closure constant</th>
<th>Remarks</th>
</tr>
</thead>
<tbody>
<tr>
<td>$C_1 = 0.302$</td>
<td>Estimated by matching the WS77 equations, in the absence of flux divergences.</td>
</tr>
<tr>
<td>$C_2 = 2.313$</td>
<td>$\mu$  = 2.4</td>
</tr>
<tr>
<td>$C_3 = 24.296$</td>
<td>$\mu$  = 1.9</td>
</tr>
<tr>
<td>$C_n = 0.099$</td>
<td>$\mu$  = 1.25</td>
</tr>
<tr>
<td>$\gamma = 0.07$</td>
<td>$C_d = 0.25$, and that the mixing length is $\gamma(z - d)$, where $\kappa = 0.4$ and $d$ is the height of zero-plane displacement.</td>
</tr>
</tbody>
</table>

Appendix 2: Sensitivity Analysis of $T_L$ profile

The results of CSO in Fig. 5 were derived by assuming a vertically constant $T_L$ profile inside the canopy with a correction for atmospheric stability. We repeated the optimization using the three plausible $T_L$ schemes described in the section ‘Velocity statistics of nocturnal conditions’ to assess how sensitive modelled $F(t,h)$ and $F(t,0)$ are to such an assumption. As shown in Table 5, based on $\frac{T_L}{T_h} = 0.3$, the model over-predicted $F(t,h)$ and $F(t,0)$ by a factor of 5 and 3, respectively. It clearly shows that such $T_L$ profile is not reasonable for nocturnal conditions. Our calculation improved significantly after $T_L$ was corrected for stability, with case 2 and 3 providing comparable results.

<table>
<thead>
<tr>
<th>Case</th>
<th>$\frac{T_L}{T_h}$</th>
<th>$F(t,h)$</th>
<th>$F(t,0)$</th>
</tr>
</thead>
<tbody>
<tr>
<td>1</td>
<td>0.3</td>
<td>5.0</td>
<td>3.0</td>
</tr>
<tr>
<td>2</td>
<td>$f_{o1}$ where $f_{o1} = \phi_1 \int [\exp(-0.095x^2 - 0.6454x - 3.8516)]$, $\phi_1 = \ln(\frac{x}{2})$</td>
<td>1.2</td>
<td>0.92</td>
</tr>
<tr>
<td>3</td>
<td>$f_{o2}$ where $f_{o2} = 0.155(z - 1)$, where $\begin{cases} x \leq 0.4, &amp; f_{o2}(x) = 0.4 \ x \geq 0.4, &amp; f_{o2}(x) = 0.155 \ \end{cases}$</td>
<td>1.37</td>
<td>0.96</td>
</tr>
</tbody>
</table>