

CHAPTER 1

AN INTRODUCTION TO THE ECONOMICS OF FOREST DISTURBANCE

Thomas P. Holmes, Jeffrey P. Prestemon, Karen L. Abt

1. FOREST DISTURBANCES AND ECONOMIC SYSTEMS

Increasing severity of recent wildfires, storms, pest outbreaks, and biological invasions has intensified concern among governmental agencies, private enterprises, and the general public regarding the future of forest resources. Economic analysis can help decision-makers understand the causes and consequences of forest disturbances, as well as evaluate trade-offs, and set priorities. It is the premise of this book that similarities existing among forest disturbances permit the development of a unified framework for economic analysis. This book sketches out how this framework could be constructed, provides an overview and summary of current research in the economics of forest disturbances, and illustrates how economic theory and empirical methods can be applied to address specific disturbances.

From an economic perspective, a forest disturbance can be defined as an event that interrupts or impedes the flow of goods and services provided by forest ecosystems that are desired by people. This definition parallels the ecological definition of a natural disturbance as "...any relatively discrete event in time that disrupts ecosystem, community, or population structure and changes resources, substrate availability, or the physical environment" (White and Pickett 1985, p. 7). Although timber harvesting and land use change are forest disturbances according to this definition, in this book we address large-scale natural disturbances that can be mediated and modified by human actions. Examples of such large-scale natural forest disturbances occurring during the past century include the chestnut blight which largely eliminated chestnut trees from hardwood forests in the eastern United States, Hurricane Katrina which blew down large swaths of forest in the United States Southeast, and the 1988 fires in Yellowstone National Park and surrounding forest ecosystems that burned more than one-quarter million hectares.

Catastrophic disturbances affect both public and private land, and the management interventions applied to mitigate damages will vary depending on the

objectives of the land manager. Management interventions are typically made with at least one of two goals in mind: (1) reducing the probability (risk) of an unwanted state of nature, and/or (2) reducing the negative consequences if an unwanted state of nature occurs. Interventions can be made prior to (prevention), during (suppression/eradication), or subsequent to (restoration/recovery) a natural disturbance event. Although interventions can be viewed as an optimal capital management problem from the perspective of a private timber manager (chapter 3), this book focuses attention on the economic analysis of interventions by public land managers and policy makers in forest disturbance processes, keeping in mind that public forest health protection programs are often designed to alter the behavior of private forest owners (chapter 19).

Although the use of economic analysis to inform decisions about the optimal level of public investment in forest protection is not new (chapters 16 and 18), the modern practice of forest disturbance economics continues to pose challenges. First, inference and prediction regarding future forest disturbances are characterized both by high variability (fluctuations not explained by deterministic processes) and uncertainty (limited knowledge of model parameters). Rigorous mathematical, statistical, and econometric models are required to address these special characteristics of forest disturbance production. Second, many of the effects of forest disturbances fall upon non-timber goods and services, such as outdoor recreation or aesthetic views. Because few non-market economic studies of disturbances have been conducted to date, and the current level of understanding of these impacts is limited, existing estimates of economic damages from forest disturbances may be severely biased by the lack of information on non-market impacts. Third, understanding the linkages between the costs of management interventions and changes in the net economic benefits provided by forest ecosystems is challenging because of the time lag between interventions and changes in the provision and value of ecosystem services. Consequently, empirical examination of the effects of protection investments requires long time spans, large data gathering efforts, and careful and innovative scientific enterprise.

To motivate subsequent analysis, the following section presents a broad characterization of various classes of forest disturbances and describes how specific disturbance characteristics constrain the set of management interventions that can be employed to mitigate economic impacts. Section 3 then provides an overview of the major lessons learned and presented throughout the remainder of the book. The chapter concludes in section 4 by offering suggestions for future research.

2. FOREST DISTURBANCE CHARACTERISTICS

Forest disturbances can be classified in three broad categories (table 1.1): abiotic events (storms, landslides, volcanoes, droughts, and floods), biotic events (insects, diseases, and invasive plants), and wildfires (a mix of abiotic and biotic disturbances). We further characterize forest disturbances using four key variables that

Table 1.1. Characteristics of forest disturbances and list of book chapters containing applied economic analysis.

Disturbance Type	Disturbance Sub-Type	Rate of Spread	Maximum Spatial Scale	Endogenous or Exogenous	Forest Management Strategy	Book Chapter
Abiotic						
Storms	Tornado	Hours	Small	Exogenous	None	—
	Hurricane	Hours → days	Very large	Exogenous	Salvage	11
	Straight-line wind	Hours → days	Large	Exogenous	Salvage	—
	Ice	Hours → days	Very large	Exogenous	Remove hazard trees	—
Volcanoes	n/a	Hours → days	Large	Exogenous	Restore	—
Floods	n/a	Hours → months	Medium	Exogenous	Restore	—
Drought	n/a	Months → Years	Very large	Exogenous	Restore	—
Biotic						
Invasive plants	n/a	Years → centuries	Very large	Exogenous	Eradicate	—
Pests	Exotic Insects and Diseases	Years → centuries	Very large	Exogenous	Eradicate; slow spread; pre-emptive harvest; bio-control; salvage	2, 3, 11, 19
	Native Insects and Diseases	Months → years	Large	Endogenous	Shorten rotation; genetic selection; chemicals; salvage	2, 3, 19
Mixed Wildfire	All Ignition Sources	Hours → months	Large	Endogenous and exogenous	Suppress; reduce fuels; salvage; restore	2-18

influence economic costs and losses: rate of spread, spatial scale, whether the source of the disturbance is endogenous (inside) or exogenous to (outside) the forest, and forest management strategies employed to mitigate impacts. Forest disturbances are of interest to economists when their frequency and size are consequential enough to induce a management or policy response, and economic analysis of forest disturbance generally seeks to identify the optimal level of intervention in disturbance processes by balancing costs and losses. Our classification scheme recognizes that the scope and type of interventions available are a function of their biotic or abiotic nature, spread rate, and sources.

2.1 Abiotic Disturbances

Abiotic disturbances are characterized as deriving from energy sources originating outside of forests, and include climatic and geologic disturbances. Abiotic disturbances are stochastic and difficult to predict far in advance. Neither the probability of occurrence nor the magnitude of effects on forests can be significantly influenced by forest management. Although manipulation of stocking density or species selection may have some effect on reducing damage from abiotic events (Wilson and Baker 2001), the main forest management strategies are to salvage timber and restore landscapes and ecosystems.

2.1.1 Climatic events

Long-term climate change, acting as a slowly changing parameter that conditions the dynamic behavior of fast moving variables, can affect the entire constellation of forest disturbance processes including fire, drought, introduced species, insect and disease outbreaks, hurricanes, windstorms, ice storms, and landslides (Dale et al. 2001). Recognizing this, here we review the fast climatic events that affect forests—tornadoes, ice storms, hurricanes, droughts, and floods.

Tornadoes may damage tens to hundreds of hectares of forest cover (Glitzenstein and Harcombe 1988, Harcombe 1988, Peterson and Pickett 1995), and therefore may have a substantial impact on individual forest owners. However, they are generally too small and infrequent to have an impact on aggregate economic welfare. In contrast, straight-line winds occasionally have large-scale, catastrophic impacts, such as the July 4, 1999, blowdown that damaged 57,000 hectares of forest in the Boundary Waters Canoe Area in northern Minnesota (Schulte and Mladenoff 2005). Other recent examples can be found in Europe (Nilsson et al. 2004). Tropical cyclones (typhoons and hurricanes) can also cause catastrophic forest damage, as was the case with Hurricane Hugo, which destroyed 1.8 million hectares of forest in South Carolina (Sheffield and Thompson 1992). Large scale climatic events can have substantial economic impacts on timber markets (chapter 9) and, additionally, can cause non-market economic losses to residential properties, public parks, and rural landscapes.

Large, damaging ice (or glaze) storms are infrequent, although they can occasionally cause tree damage over millions of hectares of urban and rural forests (Smith 2000). Management interventions typically focus on the decision of whether or not to remove damaged trees.

Major, infrequent floods can cause tree mortality if soils are saturated long enough to create anoxic conditions, which cause tree roots to die. This was the case in the Midwest flood of 1993 on the upper Mississippi River, which caused extensive mortality to trees and shrubs in the floodplain (Sparks et al. 1998).

Severe droughts can also induce economic costs and losses on forested properties, and impacts on urban forests and residential properties can be particularly severe. For example, the drought of 1934 killed about 25 percent of the trees and injured another 25 percent of the trees in Manhattan, Kansas (Stiles and Melchers 1935). The loss of aesthetic value, shading, and other non-market benefits of trees due to drought is compounded by the costs of removing and replacing dead and dying trees.

2.1.2 Geological events

Geological events are similar to climatic events, releasing large amounts of energy over a short time period. The two types of geological events that are most consequential to forests are volcanoes and landslides. Landslides occur in forested regions with steep topography and can be triggered by heavy rain or seismic events such as earthquakes or volcanoes. Earthquake caused landslides can be a major disturbance in tropical forests, and landslides ranging from 5,000-10,000 hectares have been observed (Veblen and Ashton 1978). Smaller landslides of less than 1 hectare may be quite common in tropical forests with steep slopes (Guariguata 1990). In the United States, landslides not associated with volcanoes are not known to influence forests to an economically significant extent.

Perhaps the best known volcanic eruption-related forest disturbance in the United States was the eruption of Mount St. Helens in southwest Washington in 1980. This eruption affected an area exceeding 70,000 hectares, including a variety of disturbances due to pyroclastic flow, tree blowdown, scorched trees, mudflows, and debris avalanches (Turner et al. 1997). Such occurrences in the volcanoes around the Pacific Rim are anticipated to occur every 100-1,000 years.

2.2 Biotic Disturbance

Biotic forest disturbances result from the propagation, growth, and spread of biological organisms that depend on forest resources to complete their life cycle. These disturbances include a diverse array of native and exotic insects, diseases, and invasive plants. Biotic disturbances are endogenous, and thus have a different suite of interventions available to affect the probability of occurrence and the extent of damages.

Invasive forest plants compete with native vegetation and can reduce the biological diversity of forest ecosystems. The growth and spread of invasive forest plants is relatively slow and predictable, and the primary control strategy is eradication followed by rehabilitation with fast-growing native plants (Miller 2003). A common invasive forest plant is kudzu (*Pueraria montana*), which is thought to cover 3 million hectares in the eastern United States and is expanding at the rate of 50,000 hectares per year (Forseth and Innis 2004).

Forest insects and diseases attack selected tree species, and pest outbreaks typically do not cause mortality to all trees in an infested area. However, population growth and spread can result in damages to public and private goods and services across broad landscapes. Because native trees have not co-evolved with exotic pests, they are particularly vulnerable to successful attack over the entire range of host species. Population growth of forest insects and diseases may follow non-linear or chaotic dynamics (Turchin 2003) and may be triggered or synchronized by atmospheric processes (Williams and Liebhold 2000, Liebhold et al. 2004). Insect and disease outbreaks may also interact with wildfire, complicating predictions of the timing, location, or intensity of biotic disturbances (Castello et al. 1995, McCullough et al. 1998).

The spatial spread of biotic disturbances occurs on time scales of years to centuries (e.g., gypsy moth), which is slow relative to the rate of spread of abiotic disturbances. This slower time scale, together with their host dependencies, permits a greater number of management strategies to be developed to combat biotic disturbances. Timber management strategies are based on the idea that the amount of timber at risk of damage or loss can be reduced by actions such as shortening timber rotations, pre-emptive harvesting of timber in anticipation of an imminent (actively spreading) insect or disease outbreak, and selection or propagation of trees with natural resistance to the pest (Cubbage et al. 2000). Other strategies can be used to protect the aesthetic and non-market values of trees and forests, such as tree removal, the application of chemicals to eradicate or slow the spread of insects or reduce the rate of disease progress on particular trees, and biological control. In the wake of biotic disturbances, timber salvage and ecosystem restoration strategies can be used to minimize short term economic impacts and restore long term economic values.

2.3 Wildfires

The temporal scale of wildfires is intermediate between biotic and abiotic disturbances—wildfires are briefer in duration than biotic disturbances but can be longer than abiotic disturbances. On a spatial scale, wildfires span more than four orders of magnitude (assuming that the smallest wildfire is in the order of 0.1 ha). Large wildfires can equal or exceed the size of most abiotic forest disturbances (except hurricanes) and yet are smaller in area than the most severe biological invasions.

As with biotic disturbances, wildfires are dependent on the availability of sufficient host material, and their extent and spread are limited by weather and climatic conditions. This dependency on host materials—fuels—provides the rationale for management strategies such as prescribed fire and mechanical fuel reduction which are applied with the goal of reducing wildfire spread and intensity. Because wildfires spread over hours to months, and because they often spread in relatively predictable directions, fire suppression can be used to limit fire sizes. Additionally, because the destructive character of large wildfires is patchy, substantial areas of forest may be killed while other areas remain relatively unharmed. Consequently, timber salvage following fire is often a viable management option. Restoration of areas burned by wildfires is also possible, mitigating negative impacts on watersheds and other future ecosystem values (Kent et al. 2003). Finally, it should be recognized that wildfires can convey benefits to fire dependent ecosystems, and the practice of letting some wildfires burn (referred to as “wildland fire use” in the United States) is becoming a more commonly accepted tool for public forest management (Doane et al. 2006).

3. OVERVIEW OF CHAPTERS

The structure of this book reflects our view that: (1) economic analyses of forest disturbance is enhanced by its congruence with ecological understanding (chapter 2); (2) forest disturbances can be modeled as stochastic economic production processes (section II); (3) consistent accounts of market and non-market economic effects of forest disturbances are pre-requisite to planning and decision-making (section III); and (4) economic models can be used to improve decisions taken to mitigate the negative economic consequences of forest disturbances and to set priorities (section IV). Below, we provide an overview of the contents of individual chapters.

3.1 Forest Disturbance Processes

From an economic perspective, forest disturbances are stochastic events that can be modeled as production processes. Some inputs into disturbance production are free (such as drought, lightning, or wind) and other inputs are purchased (such as capital and labor). The stochastic nature of disturbance processes suggests that disturbance outputs can be measured using probability distributions for metrics such as area burned or the number of large fires (chapters 2-7). By conducting statistical and econometric analyses, the economic consequences of management interventions can be identified as shifts in the stochastic distribution of disturbance events that occur in response to the application of purchased inputs.

Forest disturbances are characterized by high variance in the scale of physical and economic impacts (chapters 2-5) which can be explained by a number of factors. First, favorable site conditions for disturbance establishment and spread

vary irregularly over time and space. Second, prior disturbances condition the landscape for subsequent events. Third, stochastic exogenous factors such as weather strongly influence the size of individual forest disturbances (chapters 4-6). Fourth, disturbances may be highly nonlinear in their responses to managerial and free inputs, resulting in discontinuous and catastrophic ecosystem behavior (chapter 2).

The processes that govern forest disturbances also include human caused wildfire via unintentional (e.g., campfires and debris burning escapes) and intentional behavior (arson). Arson wildfires can be understood as a production process involving a combination of weather and climate-dependent fuel conditions, economic variables, penalties, and psycho-social phenomena (chapter 7). Consequently, law enforcement and public education campaigns may be effective at reducing the frequency of arson and accidental fires. Managers may be able to mitigate the impacts of arson and other human caused fires through fuels management and pre-positioning of suppression resources.

3.2 Valuing the Economic Impacts of Forest Disturbances

The perspective presented in this book is that a full accounting of the costs and economic losses due to forest disturbances is prerequisite to effective planning and priority setting. The first step is to establish a consistent accounting and data collection framework (chapter 8). Economic systems are connected over time and space—many goods and services are substitutes and complements in consumption, and many inputs are substitutes and complements in production—and economic assessments are sensitive to spatial scale (geographic area to be assessed), temporal scale (time span used to assess impacts), and sectoral scale (economic sectors to be included). Economic assessments need to be conducted across multiple scales, and decision-makers need to be informed of the sensitivity of economic measures to the scale at which economic models are applied.

Forest disturbances such as insect epidemics, hurricanes, and wildfires can have extreme impacts on markets for goods obtained from forests. In timber markets, timber losses and damages affect economic equilibria, both through the pulse of timber salvaged from an event and through reductions in stocks of standing timber (chapter 9). Economic welfare is redistributed after a catastrophic forest disturbance, with some economic agents gaining (e.g., consumers of wood products in the short-term) while others lose (e.g., producers of damaged timber). Timber salvage policies instituted by governmental agencies should be sensitive to the redistributive impacts of such policies.

Forest disturbances can induce a significant loss of wealth for private property owners in the wildland-urban interface. For example, changes in risk perceptions resulting from nearby catastrophic wildfires can induce private property value losses reaching millions of dollars in a single community (chapter 11). Similarly, tree mortality caused by an exotic forest insect can cause losses in property

values exceeding a million dollars in an individual community due to the loss of aesthetic values and the costs associated with tree removal (chapter 11). Because wildfires reduce the value of private residential properties, private homeowners have a substantial willingness to pay for public programs designed to protect residences and communities from wildfires (chapter 12).

Wildfires can destroy recreational infrastructure and can alter the quality of outdoor recreation sites. Although few studies have been conducted to evaluate the impact of wildfires on the demand for outdoor recreation, preliminary evidence suggests that wildfires may increase the number of Wilderness visitors in the short-run, due to an influx of curiosity-seekers (chapter 10). Over the span of several decades, however, the economic value of wilderness areas that have experienced large wildfires may decrease because of visitation reductions brought about by the loss of mature forests and the presence of less desirable forest conditions. More research is needed to understand the impacts of wildfires, storms, and invasive species on all forms of outdoor recreation.

3.3 Decision Making in Response to Forest Disturbances

Forest disturbances typically involve an element of surprise, and forest protection decisions must be made before the ultimate state of nature is revealed. A general approach to forest protection is to reduce the risk (probability) that an unwanted state of nature will occur and to take steps that would reduce negative consequences in the event that an unwanted state does in fact occur (chapter 19). One example of this approach is evidenced by the various state and local governmental agencies that have established programs to reduce wildfire hazards in high risk areas through regulations on land-uses and vegetation management (chapter 14). Another example is provided by fuel management programs implemented by private and public forest landowners, which have been shown to reduce both damages and subsequent suppression costs (chapter 13). Much may be learned by examining the successes and shortcomings of existing programs and policies.

Another approach to managing uncertainty about future conditions is to construct forecast models using the best available data. Econometric forecasts of future wildfire suppression costs provide a rigorous means of establishing budget requests by federal land management agencies (chapter 17). Econometric models can also quantify the degree of uncertainty about parameter values and test hypotheses about proposed driving variables. Loss functions can be used to compare the performance of various models and allow managers to use planning tools in ways that reflect their priorities and risk perceptions.

Economists are cognizant of the role that incentives play in decision-making (chapter 16). Incentives regarding wildfire suppression and overall fire program management influence the costs and benefits of high profile suppression efforts by federal agencies. For example, funding wildfire suppression with emergency funds provides little incentive for cost containment (chapter 16). Further, because

wildfires can produce ecological benefits, recognition by incident managers of the fuel treatment or other benefits of fire could facilitate improvement of management approaches and reduce associated costs (chapters 15 and 16).

Programs designed to protect forest ecosystems are complex and include many interacting components. For example, governmental programs for fire management include components for fire prevention, detection, fuels management, suppression, and post-fire site rehabilitation. Because of the linkages and feedbacks between components, economic efficiency is compromised when analysis is conducted component-by-component (chapter 18). The development of integrated forest protection programs will likely be worthwhile but present significant challenges because they require models and tools that accurately describe the trade-offs among alternative program inputs.

Forest health protection from invasive species is a public good, in that the benefits from forest protection are shared by other members of the community. This context provides the justification for government intervention. Further, forest health protection is a weakest-link public good. The weakest-link character of forest health protection relegates the level of forest protection attained by a community to the weakest members of the community. Consequently, effective forest health protection programs require that the weakest links be strengthened by targeting information to those most likely to engage in risky behavior. In particular, information describing the weakest-link nature of forest protection should be targeted at private landowners to enhance the likelihood that they will participate in forest protection programs (chapter 19). Weakest links can be identified using economic surveys of household behavior.

4. RESEARCH NEEDS

Economic models need to account for the complexity of disturbance processes so that the efficiency and efficacy of management interventions can be realistically assessed. Nonlinear dynamics and spatial diffusion are challenging attributes of forest disturbances, and further development of statistical, econometric, mathematical, and simulation models that address management interventions across various temporal and spatial scales are needed. In particular, we suggest that research is needed that would enhance the ability to predict catastrophic changes in ecological and economic variables.

Preliminary evidence suggests that the non-market economic impacts of forest disturbances are substantial, but few studies have been conducted. Further studies of the economic damages caused by forest disturbances to private property values, to ecosystem service values provided by public and private forests, and to human health (e.g., smoke from wildfires, wildland use fires, and prescribed fires; dermatitis from caterpillars) are needed. A more comprehensive understanding of non-market economic impacts would illuminate the severity of these threats and provide a larger knowledge base for improved management decisions.

Fire and forest health protection programs need to be evaluated as integrated systems, rather than being evaluated in isolation. The wildfire program is an example, where analysis has focused on the market effects of timber damages from wildfire and wildfire suppression costs. Yet wildfire programs also encompass the outcomes from fuels management and the potential positive impacts of restoring ecosystem function and reducing future wildfire. More generally, forest disturbances such as wildfire, insect and disease outbreaks, and biological invasions interact across broad spatial and temporal scales. Economic and ecological models for integrating the various components of fire and forest health protection programs are needed and will likely lead to lower program costs and greater benefits to society.

The time lag between the imposition of a management intervention and the occurrence of a catastrophic event creates uncertainty about the efficacy of management actions. Models of decision-making under uncertainty is a key topic for future research, and models that incorporate learning as new information is revealed are needed.

Finally, data required for economic analyses of forest disturbances still need improvement. Although economists have developed specialized econometric methods for analyzing non-experimental data, the data available for analyzing forest disturbances is often inconsistent, fragmentary, or unavailable over the time spans at which disturbance processes operate. Improved coordination between economists and the data collection operations conducted within land management agencies would enhance the ability for economists to evaluate trade-offs and provide meaningful and timely information to policy-makers.

5. REFERENCES

- Castello, J.D., D.J. Leopold, and P.J. Smallidge. 1995. Pathogens, patterns, and processes in forest ecosystems. *Bioscience* 45(1):16-24.
- Cubbage, F.W., J.M. Pye, T.P. Holmes, and J.E. Wagner. 2000. An economic evaluation of fusiform rust protection research. *Southern Journal of Applied Forestry* 24(2):77-85.
- Dale, V.H., L.A. Joyce, S. McNulty, R.P. Neilson, M.P. Ayres, M.D. Flannigan, P.J. Hanson, L.C. Irland, A.E. Lugo, C.J. Peterson, D. Simberloff, F.J. Swanson, B.J. Stocks, and B.M. Wotton. 2001. Climate change and forest disturbances. *Bioscience* 51(9):723-734.
- Doane, D., J. O'Laughlin, P. Morgan, and C. Miller. 2006. Barriers to wildland fire use: A preliminary problem analysis. *International Journal of Wilderness* 12(1):36-38.
- Forseth, Jr., I.N., and A.F. Innis. 2004. Kudzu (*Pueraria montana*): History, physiology, and ecology combine to make a major ecosystem threat. *Critical Reviews in Plant Sciences* 23(5):401-413.
- Glitzenstein, J.S., and P.A. Harcombe. 1988. Effects of the December 1983 tornado on forest vegetation of the Big Thicket, Southeast Texas, U.S.A. *Forest Ecology and Management* 25:269-290.

- Guariguata, M.R. 1990. Landslide disturbance and forest regeneration in the upper Luquillo Mountains of Puerto Rico. *The Journal of Ecology* 78(3):814-832.
- Kent, B., K. Gebert, S. McCaffrey, W. Martin, D. Calkin, I. Schuster, H. Martin, W. Bender, G. Alward, K. Yoshitaka, P.J. Cohn, M. Carroll, D. Williams, and C. Ekarius. 2003. Social and economic issues of the Hayman Fire. In: Hayman Fire Case Study, R.T. Graham (ed.). Gen. Tech. Rep. RMRS-114. USDA Forest Service, Fort Collins, CO.
- Liebholt, A., W.D. Koenig, and O.N. Bjørnstad. 2004. Spatial synchrony in population dynamics. *Annual Review of Ecology, Evolution, and Systematics* 35:467-490.
- McCullough, D.G., R.A. Werner, and D. Neuman. 1998. Fire and insects in northern and boreal forest ecosystems of North America. *Annual Review of Entomology* 43:107-127.
- Miller, J.H. 2003. Nonnative invasive plants of southern forests: A field guide for identification and control. Gen. Tech. Rep. SRS-62. USDA Forest Service, Asheville, NC.
- Nillson, C., I. Stjernquist, L. Barring, P. Schlyter, A.M. Jönsson, and H. Samuelsson. 2004. Recorded storm damage in Swedish forests 1901-2000. *Forest Ecology and Management* 199:165-173.
- Peterson, C.J., and S.T.A. Pickett. 1995. Forest reorganization: A case study in an old-growth forest catastrophic blowdown. *Ecology* 76(3):763-774.
- Schulte, L.A., and D.J. Mladenoff. 2005. Severe wind and fire regimes in northern forests: Historical variability at the regional scale. *Ecology* 86(2):431-445.
- Sheffield, R.M., and M.T. Thompson. 1992. Hurricane Hugo: Effects on South Carolina's forest resource. Research Paper SE-284, USDA Forest Service, Asheville, NC.
- Smith, W.H. 2000. Ice and forest health. *Northern Journal of Applied Forestry* 17(1): 16-19.
- Sparks, R.E., J.C. Nelson, and Y. Yin. 1998. Naturalization of the flood regime in regulated rivers. *Bioscience* 48(9):706-720.
- Stiles, E.H., and L.E. Melchers. 1935. The drought of 1934 and its effect on trees in Kansas. *Transactions of the Kansas Academy of Science* 38:107-127.
- Turchin, P. 2003. *Complex population dynamics: A theoretical/empirical synthesis*. Princeton University Press, Princeton, NJ.
- Turner, M.G., V.H. Dale, and E.H. Everham, III. 1997. Fires, hurricanes, and volcanoes: Comparing large disturbances. *BioScience* 47(11):758-768.
- Veblen, T.T., and D.H. Ashton. 1978. Catastrophic influences on the vegetation of the Valdivian Andes, Chile. *Vegetatio* 36(3):149-167.
- White, P.S., and S.T.A. Pickett. 1985. Natural disturbance and patch dynamics: An introduction. In: *The Ecology of Natural Disturbance and Patch Dynamics*, S.T.A. Pickett and P.S. White (eds.). Academic Press, Inc., New York, NY.
- Williams, D.W., and A.M. Liebhold. 2000. Spatial synchrony of spruce budworm outbreaks in eastern North America. *Ecology* 81(10):2753-2766.
- Wilson, J.S., and P.J. Baker. 2001. Flexibility in forest management: Managing uncertainty in Douglas-fir forests of the Pacific Northwest. *Forest Ecology and Management* 145:219-227.