DECAY OF FIRE-CAUSED SNAGS IN OCALA SAND PINE

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ABSTRACT

Sand pine (*Pinus clausa*) scrub is adapted to, and regenerated by, periodic stand-replacement wildfire, which consumes the understory and kills the overstory. The heat of the fire opens the serotinous cones of Ocala sand pine (*P. clausa* var. *clausa*), releasing quantities of seed that reestablish the overstory, while the understory regenerates by sprouting or from soil-stored seed. Resource managers attempt to mimic this process by clear cutting and direct seeding, which seems to work quite well. However, this method results in far fewer snags, which may have important functions in the sand pine scrub ecosystem. Because sand pine is a short-lived species with very little heartwood, it is believed that these snags would be a temporary feature of the system, and thus were not that critical. My study tests this assumption by documenting the life of snags following a stand-replacement fire in sand pine scrub on the Ocala National Forest, Florida. Three stands were burned, one by a prescribed burn in May 1993 and two by natural fire in August 1993. Prior to the fire, there were 520 Ocala sand pine/ha, 96 oaks/ha, and 137 snags/ha. The fire killed all of the sand pine and most of the oaks. Decay proceeded more slowly than expected in the resulting snags. After 2 years 69% retained most of the bark, 27% had lost all bark and most of the limbs, and only 4% had visible sapwood decay. The first snags fell between 12 and 18 months following the fire, but it took 5 years for 50% of the snags to fall. At the end of 7 years following the fire, 32% of the sand pine snags were still standing. Thus, the snags in sand pine scrub occur as part of the structure of the sand pine ecosystem for much longer than expected. Managers may need to consider more prescribed fire in wilderness areas to generate these snags.

keywords: coarse woody debris, ecosystem management, fire in ecosystem management, Florida, Ocala National Forest, *Pinus clausa*, sand pine, snags.

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INTRODUCTION

Sand pine scrub grows on deep droughty infertile sands of marine and aeolian origin. Water and wind formed these features as sea levels fluctuated during past glacial and interglacial periods (Kurz 1942, Laessle 1958, Snedaker et al. 1972). Because of washing and sorting during transport and deposition, soil parent material was nearly pure quartz sand (Laessle 1958). This produced soils which are almost exclusively entisols and mostly Quartzipsamments (Myers 1990) typified by the Astatula, Lakeland, Paola, and St. Lucie soil series.

The Ocala variety of sand pine is native to scrubs on the central ridge of Florida and a strip of old dunes stretching from St. John's County south to the northern portion of Dade County (now Miami-Dade County) on the east coast and from near Cedar Key south to Naples on the west coast (Small 1921, Harper 1927, Myers 1990). Choctawhatchee sand pine (*P. clausa* var. *immuginata*) is the dominant tree in scrubs along the Gulf Coast, including off-shore islands, of northwest Florida from the Apalachicola River westward into Alabama. Interior scrub, the largest contiguous area of Ocala sand pine scrub, occupies about 101,000 ha on the Ocala National Forest (Brendemuehl 1990). This area has hot, humid summers, somewhat dry winters, and a long growing season averaging 300 days. Precipitation is abundant averaging 1300 mm per year (1960–1980) with a maximum in July and a minimum in May. Even though rainfall is plentiful, the area experiences frequent drought periods, which can develop within 2 weeks of a heavy rainfall because of the low soil moisture-holding capacity.

Sand pine scrub is a unique community with a suite of species that occur nowhere else (Christman and Judd 1990). Ocala sand pine scrub has an overstory of even-aged sand pine with twisted and leaning trunks over an understory of evergreen shrubs (Myers 1990). Typical understory species include myrtle oak (*Quercus myrtifolia*), sand live oak (*Q. geminata*), Chapman's oak (*Q. chapmanii*), turkey oak (*Q. laevis*), rusty lyonia (*Lyonia ferruginea*), rosemary (*Ceratiola ericoides*), scrub palmetto (*Sabal etonia*), and saw palmetto (*Serenoa repens*) (Outcalt 1997). Because of the well-drained soils and competition from the sand pine overstory and understory shrubs, herbs and grasses are very sparse in mature scrub habitats. Typical species include beak rush (*Rhynchospora megalocarpa*), milk peas (*Galactia* spp.), and bluestems (*Andropogon* spp.). Lichens (*Cladonia* spp.) form extensive patches on the forest floor (Greenberg et al. 1995*b*).

Species richness and diversity of herbaceous plants are significantly greater, however, in recently disturbed areas (Greenberg et al. 1995*b*). Most of the endemic species thrive in this open scrub where sand pine is in the seedling stage. As with many ecosystems in the southeastern U.S., the primary disturbance creating these open conditions is fire. Because of its sparse ground cover and compacted litter layer, much of the time sand pine scrub is virtually fireproof. However, every 10 to 100 years, usually during the spring

Table 1. Characteristics (mean ± 95% confidence interval) of overstory trees prior to prescribed fire in 1993 in Ocala sand pine scrub on the Ocala National Forest, Florida.

Characteristic	Stand 1	Stand 2	Stand 3
Ocala sand pine			
Density (no./ha)	670 ± 137	470 ± 78	367 ± 131
Diameter (cm)	17.5 ± 0.6	20.1 ± 0.9	20.3 ± 1.2
Basal area (m^2/ha)	16.8 ± 1.6	13.8 ± 4.0	10.0 ± 3.5
Height (m)	16.9 ± 0.8		
Oaks			
Density (no./ha)	77 ± 48	157 ± 64	53 ± 39
Diameter (cm)	6.6 ± 0.8	7.0 ± 0.6	8.4 ± 0.7
Height (m)	4.2 ± 0.4		

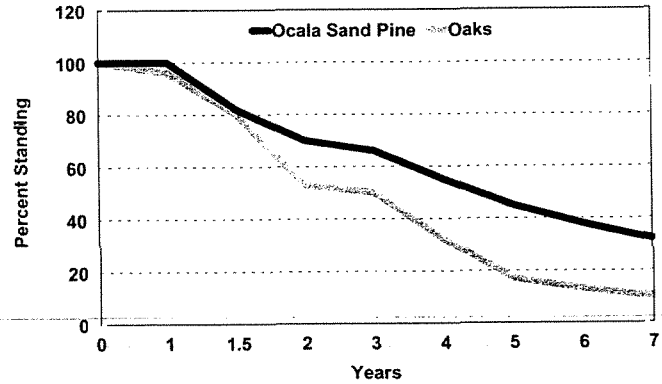

Fig. 2. Percentage of Ocala sand pine and scrub oak snags standing, following prescribed fire in 1993 in sand pine scrub on the Ocala National Forest, Florida.

at least 10 cm less than the sand pine (Table 1). These oaks composed the midstory layer of the stands as shown by the much smaller average height.

The fires scorched most needles on the sand pine trees, but some trees along the edge retained green crowns. Two months following the fires, all pines in the stands had brown needles and were obviously dead. During the first year, tops and branches were lost from a few of the dead sand pines. Decay was rather slow, however, with just 4% of the sand pine snags losing most or all of their bark during the first year. The first stems fell over during the second season following the fire (Figure 2). Ocala sand pine stems continued to fall the following years but it took over 4 years before 50% of them were on the ground. After 7 years, about one-third of the stems remained standing. Midstory scrub oaks in the stands began falling sooner and the 50% loss figure was reached a year sooner, at 3 years. Nearly all of the oak stems had fallen after 7 years.

Diameter of sand pine stems had little effect on snag longevity, as the rate of loss was essentially the same for a 6-cm stem as it was for a 26-cm stem (Figure 3). The loss of Ocala sand pine snags was relatively constant, declining about 8.5% per year. Assuming the decay rates remain constant, all snags will likely have fallen by 11 years following the fires.

Although diameter did not affect rate of snag loss, longevity did seem related to fire intensity as shown by data from stand 1. Snags separated out into two general groups with those in the higher-intensity categories falling over more quickly than those less intensely burned (Figure 4). Those sand pine stems where fire was intense enough to melt 75% or more of the metal tag had a higher initial loss rate and declined to the 50% snag loss level by 3 years following the fires. The less intensely burned trees did not reach the 50% snag loss level until more than 5 years following the fire.

The progression of decay can be tracked by following the number of pine stems in each category. After 1.5 years, most of the pine snags still had much of their bark (Figure 5). Two years after the fires, there was a shift to more stems with some or most of the bark gone. This continued through years 3 and 4 along with an increase in the number of fallen stems. After 5 years, most of the stems had lost their bark and by 6 years, many had significant sapwood decay. Seven years following the fires there were no stems with the bark still on and nearly all had at least some sapwood decay. Few stems, however, had become completely soft and none had yet begun to lose shape. Scrub oak snags retained their bark longer than the sand pine with 66% retaining all bark 2 years following the fires. Sapwood decay was more rapid, however, once it be-

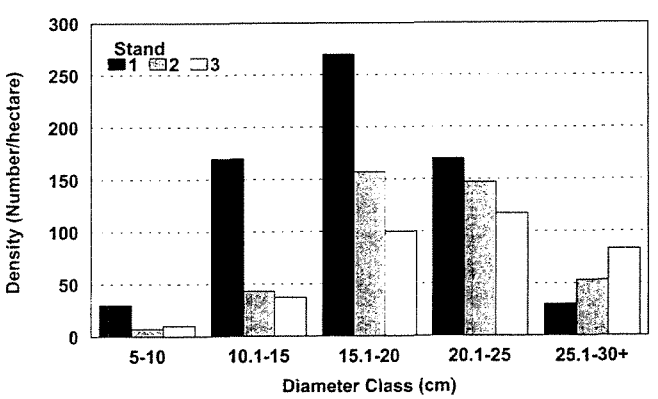

Fig. 1. Density of Ocala sand pine by diameter class, prior to prescribed fire in 1993, in sand pine scrub on the Ocala National Forest, Florida.

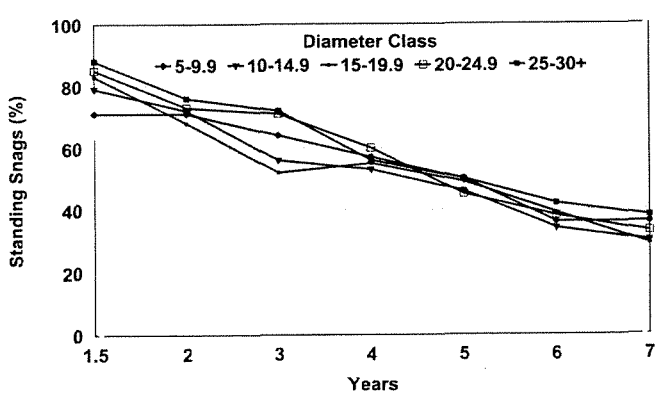

Fig. 3. Effect of diameter class on percentage of Ocala sand pine snags standing following prescribed fire in sand pine scrub on the Ocala National Forest, Florida.

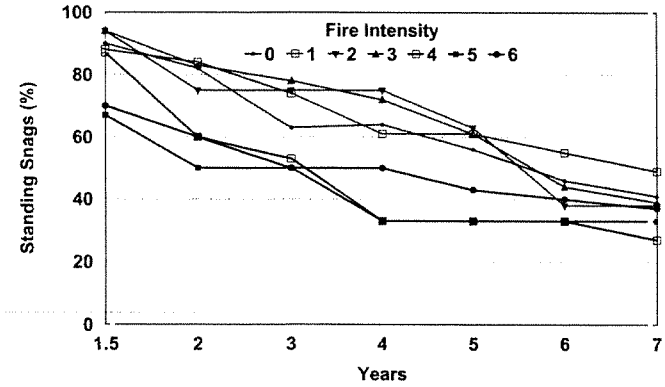

Fig. 4. Effect of fire intensity in stand 1 on percentage of Ocala sand pine snags standing over time following prescribed fire in 1993 in sand pine scrub on the Ocala National Forest, Florida.

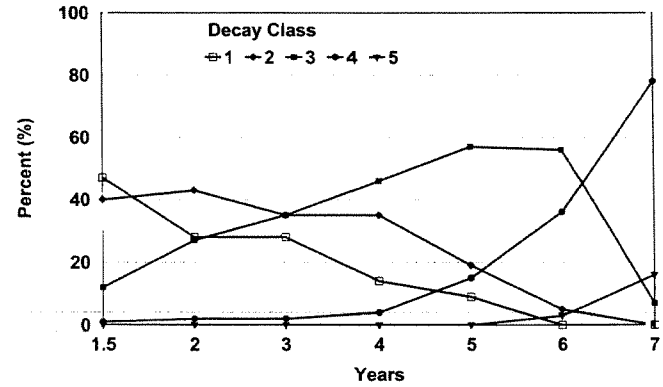

Fig. 5. Progression of decay in snags and fallen logs following prescribed fire in 1993 in sand pine scrub on the Ocala National Forest, Florida.

gan with 20% of the oak stems losing shape and 10% were just fragmented piles of wood 7 years after the fires.

DISCUSSION

Because Ocala sand pine has little heartwood, it has been presumed it would decay quite rapidly and that standing snags would be a temporary structural component of burned stands. Once Ocala sand pine dies, the first stage of change is the loss of limbs and crowns, which began during the first 6 months following the fires. This is followed by the loss of bark from the bole, which began at 1.5 years and continued until nearly all stems had lost their bark at 5 years following the fires. It took 6 years, however, for most of the pine snags to develop significant sapwood decay. Because this sapwood decay proceeded more slowly than anticipated, snags lasted much longer than expected, with none falling until 1.5 years after the fires and 50% still standing after 4 years. Even after 7 years one-third of the snags were still standing, although many were broken and less than 6 m tall. Based on the rate of loss to date, it should take over 10 years for all of the snags to fall. This loss rate is only slightly faster than reported for fire-killed lodgepole pine (*Pinus contorta*) snags in Montana (Lyon 1977). It also followed a similar pattern to that found for the lodgepole pine snags that had few losses the first 2 years and then a steady decline thereafter.

The scrub oak snags were less durable than the pine: they began falling sooner and continued at a greater rate. Only a few were still standing after 7 years. In a study in the Southern Appalachians, Harmon (1982) also reported that oak snags decayed more rapidly than pine. Although loss rates were higher than for pines they were not nearly as rapid as those reported by Cain (1996) for herbicide-killed trees. His equation for similar-sized hardwoods (under 11 cm in diameter) predicts only 20% of the scrub oaks still standing after 2 years, while the actual amount was 52%. This could be due to a number of factors including cause of mortality, climate, or species differences. Conner et al. (1983) reported that herbicide-killed oaks decayed more rapidly than did girdled oak trees in an east Texas study. Thus, cause of death seems the most probable cause of the lower loss rate.

Raphael and Morrison (1987) found small-diameter snags of Jeffrey pine (*Pinus jeffreyi*), red fir (*Abies magnifica*), and white fir (*Abies concolor*) in California had a higher rate of loss than larger stems. Other studies have also reported smaller-diameter hardwood snags tend to fall more quickly than larger snags (Dickson et al. 1983, Cain 1996). However, tree diameter did not influence the rate of pine snag fall in my study. This is likely due to differences in species and diameter ranges. Western trees are much larger than sand pine and thus the study by Raphael and Morrison (1987) included snags with diameters exceeding 50 cm. Below 38 cm snags had an equal loss rate and it was only the larger snags which were more persistent. Thus, their data agree with my findings across the range of sand pine diameters (6 to 35 cm). There are likely species differences also, as lodgepole pine, a tree with similar ecology and physiological attributes to Ocala sand pine, showed no relationship between diameter and rate of falling (Raphael and Morrison 1987) even for the largest snags.

MANAGEMENT IMPLICATIONS

Although fire-killed snags in short-lived species like Ocala sand pine and lodgepole pine are not as durable as Douglas-fir (*Pseudotsuga menziesii*) snags, which can last for decades (Cline et al. 1980), they may be just as important in their respective ecosystems. Ocala sand pine snags are an important structural component for at least a decade following fires. It appears that cooler fires might increase the longevity of snags even further, as those burned less intensely stood longer in this study. Since the fire return interval is every 10 to 100 years, this means under natural fire regimes Ocala sand pine snags occupied a given area at least 10% of the time or on a landscape scale would occur on at least 10% of the area at any given time. In addition snags will continue to be important as coarse woody debris on the forest floor for some years after they fall as most have not yet reached the stage

where all sapwood is soft and the log is beginning to lose shape 7 years after death. These snags and fallen logs represent a significant habitat that species have evolved to exploit. Current management prescriptions call for leaving 4 to 6 snags per hectare based on needs of cavity-nesting birds (McComb et al. 1986). This is supplying snags, but it is a much different structure than would have occurred in natural regimes. Thus, managers may need to consider more prescribed fire in wilderness areas to generate the higher-density snag habitats. They also need to evaluate the function of the snags on areas burned by wildfire before they decide to salvage them for the revenue they can generate. These snags may have a much higher value to the ecosystem than previously believed.

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